

Scavenging Dynamics of the Australian Alps



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This is to certify that to the best of my knowledge; the content of this thesis is my own work.

This thesis has not been submitted for any degree or other purposes.

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1 Abstract

2 Carrion (dead animal biomass) has received little attention as a high energy and nutrient rich
3 ecosystem resource. Recently, the ability of carrion to attract diverse groups of vertebrate
4 scavengers has been utilised as a focal point within ecosystems to assess scavenging dynamics.
5 Fluctuations in carrion biomass and differences in vertebrate scavenging rates are often linked
6 to the seasons. However, despite the strong consensus that scavenging dynamics are seasonal,
7 many field-based studies still fail to sufficiently consider the seasons and/or or replicate studies
8 across all seasons. To address these shortcomings and highlight the importance of seasonal
9 effects in scavenging ecology, this thesis describes vertebrate scavenging dynamics in the
10 highly seasonal Australian Alps. Here, 15 carcasses were experimentally deployed each season
11 to monitor vertebrate scavenging dynamics for 60 days via a remote camera. The findings
12 demonstrated highly seasonal scavenging dynamics but were unique given that seasonal
13 scavenging trends were dominated by highly abundant low-ranking mesoscavengers. The
14 brushtail possum (*Trichosurus vulpecula*) dominated winter-time scavenging, presumably to
15 supplement a lack of other available food sources. Whilst raven species (*Corvus coronoides*
16 and *Corvus mellori*) were highly prolific scavengers during spring, likely to meet the demands
17 of breeding. This mesoscavenger trend was apparent despite the presence of an apex predator,
18 the dingo (*Canis dingo*). However, the dingo was found to scavenge at low rates, and this raised
19 several questions regarding density dependent predator theories; carrion abundance, and
20 availability; and how these potentially limit the top-down scavenging effects of the dingo. In
21 conclusion, the methods used here serve as a practical example of a robust and repeatable
22 experimental design for monitoring and assessing scavenging dynamics and exemplifies the
23 inclusion and importance of seasonal effects in scavenging ecology.

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38 **Chapter 1 – Introduction**

39 **Carrion in the landscape**

40 The cycle of life and death is one of the few predictable processes on Earth. Within an
41 ecological framework, this cycle provides a means to an end by recycling the energy, nutrients,
42 and resources, accumulated by an organism over its lifetime, into the broader environment as
43 it decomposes [1, 2]. Until recently, the sum of dead matter contributing to this cycle was
44 treated as a single resource pool [3]. This resource pool is dominated by plant-based detritus
45 given that its living counterparts total global biomass (450Gt) shadows that of any other
46 biomass types: bacteria (70Gt); fungi (12Gt); and animal (2Gt) [1, 4]. Because this interface
47 between life and death is dominated by the sheer scale in which plant biomass exists, the
48 subsequent ecosystem processes (e.g., trophic structuring and ecological cascades) serviced by
49 the decomposition of plant-based detritus are well understood [3]. Consequently, a significant
50 bias exists in the literature whereby the ecosystem services provided by the decomposition of
51 non-plant-based detritus are likely underestimated [1, 5, 6].

52 Carrion (dead animal biomass) has received little attention in the literature as an available
53 resource within ecosystems. Compared to plant-based detritus carrion is a much more spatially
54 and temporally patchy resource [3, 7]. In highly seasonal environments, for instance,
55 fluctuations in carrion biomass can be extreme due to processes interlinked with the seasons
56 such as thermal extremes, food availability, and breeding seasons [8]. For example, each year
57 *en masse* Pacific salmon (*Oncorhynchus* spp.) migrate from the ocean to coastal feeding
58 freshwater rivers and creeks to spawn. This salmon run only happens over a 2–3-month period
59 after which all the salmon succumb to a rapid senescent death. Gende et al. (2004) found that
60 in one Alaskan creek alone, the total run size for a season was 674 salmon. This run size equated
61 to this single creek receiving 80.2 kg of N, 11.6 kg of P, and 1.2×10^7 kJ of energy in the form

62 of carrion [9]. The fate of salmon carcasses (i.e., scavenged or *in situ* decomposition) also
63 determines ecosystem specific flow-on effects [10, 11]. For example, nutrients and energy
64 derived from salmon carcasses scavenged by bears often enter terrestrial systems and
65 accumulate in vegetation after being carried ashore, partially consumed, and left to decay, or
66 via bear urine and faeces [12, 13]. Conversely, nutrients and energy derived from carcasses
67 that come to rest within the waterway, enter the aquatic system and can supplement lower
68 trophic level species (e.g., algae) that sustain such aquatic ecosystems [14].

69 Anthropogenic impacts can also increase carrion loads in ecosystems. Roads pose a significant
70 threat to a vast array of wildlife and as such have become one of the most common places
71 where carrion can be observed [15, 16]. For example, in the United States alone it has been
72 estimated that between 89-340 million birds die annually as roadkill [17] and that vehicle
73 collision now accounts for more vertebrate mortality than hunting [18]. In Australia, it is
74 estimated that marsupial roadkill exceeds 4 million individuals per year [19]. There is also
75 evidence that mass animal mortality events are on the rise globally because of anthropogenic
76 factors including climate change [20]. These events can similarly result in large carrion inputs
77 into ecosystems. The historical tendency to overlook carrion as a resource, when considered in
78 conjunction with the increasing rates with which anthropogenic impacts are altering the carrion
79 pool, highlights the need to further understand the role of carrion within ecosystems.

80 **Carrion as a resource**

81 Despite carrion being much more spatially, temporally, and quantitatively variable than plant-
82 based detritus, its rate of decomposition is in fact 10-100 times faster than that of plant-based
83 detritus [1, 21]. This is because carrion is a much more energy and nutrient rich resource
84 characterised by higher nitrogen, phosphorus, and water contents, and fewer deterrent
85 compounds [7, 22]. Consequently, carrion is a valuable high-quality resource to which a unique
86 group of species that scavenge have become specifically adapted to exploit. The ecosystem

87 services provided by scavengers are often overlooked despite the critical role they play within
88 all ecosystems, benefiting both the natural and human environment [23]. This is especially true
89 given the recognition that predators receive for the ways in which their ecosystem services
90 structure communities across multiple trophic levels [24, 25]. Scavengers have evolved to
91 become a highly specialised group of species that in unison can efficiently break down and
92 consume carrion [26]. This fast rate at which carrion is processed makes it difficult to observe
93 the inner workings, intricacies, and dynamics of scavenging in a natural setting. However,
94 scavenging has been demonstrated to (i) provide critical linkages in food webs; (ii) distribute
95 nutrients within and among ecosystems; and (iii) inadvertently reap economic and human
96 health benefits related to carcass disposal and sanitary measures [27].

97 **Scavenger guilds and their members**

98 The group of taxa responsible for scavenging includes both vertebrate and invertebrate species
99 as well as bacteria/microbes, all of which are globally prevalent in both terrestrial and aquatic
100 ecosystems. For the purposes of this thesis, discussion will be primarily focusing on the
101 vertebrate scavengers. Broadly, vertebrate scavenger guilds are composed of obligate
102 scavengers and facultative scavengers. The former relies entirely on carrion as a food resource,
103 and it is this specificity that makes them so rare. As such, vultures are the only true terrestrial
104 vertebrate obligate scavengers due to their ability to soar over large areas in search of carrion
105 which is more or less spatially and temporally consistent within their landscapes [8, 28]. The
106 social behaviours of vultures, the primary mechanism of which is local enhancement, also
107 further facilitates their ability to locate carcasses [29, 30]. However, due to widespread human
108 persecution, many vulture species are experiencing catastrophic population declines which in
109 turn is altering scavenging dynamics globally [31].

110

111 **Facultative scavengers**

112 All other terrestrial vertebrate scavenging species are facultative scavengers. These species are
113 not reliant on carrion as their primary food resource. Instead, they scavenge on carrion at
114 different points in their life stage, in response to seasonal changes, in low resource times, or in
115 the absence of competitively dominant scavengers [32, 33]. Of the facultative scavengers,
116 predators (that scavenge) are some of the most common species recorded at carcasses. These
117 predators are typically classified as either apex predators or mesopredators. The former are
118 species characterised by their position at the top of the food chain and lack natural predators
119 [34], whilst the latter are any “midranking predator in a food web, regardless of size or
120 taxonomy” [35]. The presence or absence of either species group within ecosystems can cause
121 trophic cascades that can result from their interspecific interactions or lack thereof [34]. One
122 of the most well-known of these cascades is described by the mesopredator release hypothesis.
123 This phenomenon explains how constraints to the population growth of mesopredators, which
124 are controlled by competitively dominant apex predators, are released following a decline in
125 apex predator population levels [25, 36, 37]. The consequences of such interactions can also
126 have profound effects on scavenging dynamics when occurring surrounding carrion i.e.,
127 ‘mesoscavenger release’ [38, 39].

128 Apex predators can also influence scavenging dynamics by way of other top-down effects and
129 this can come via two main pathways [40]. *Firstly*, the predatory activities of apex predators
130 can dictate scavenging rates within an ecosystem via the provision of carrion from their
131 partially consumed prey [41]. This dynamic has been demonstrated in Yellowstone National
132 Park following the reintroduction of the grey wolf (*Canis lupus*). Prior to the reintroduction,
133 carrion biomass pulsed in March-April when many elk (*Cervus canadensis*) succumb to the
134 harsh conditions of winter. Consequently, much of the scavenging activity in Yellowstone
135 National Park primarily occurred during this period. However, carrion biomass is now more

136 seasonally available via the remains left from grey wolf kills, and thus, scavenging rates have
137 followed a similar trend [42].

138 *Secondly*, apex predators can dictate scavenging dynamics via their own scavenging activities
139 and interspecific interactions surrounding non-prey killed carrion [43]. Apex predators, like
140 vultures, can rapidly consume carrion biomass including bones, and their presence at carcass
141 sites, can therefore, accelerate decomposition rates [44]. Through fear effects (i.e., smaller
142 species avoiding larger species), the scavenging activities of apex predators can also reduce
143 scavenger species richness and the time spent scavenging by other scavenger species at carrion
144 [39, 45, 46]. Furthermore, kleptoparasitic scavenging (i.e., scavenging of stolen prey) by apex
145 predators can have compounding effects on the victim predator whose kill has been stolen. Not
146 only will the victim predator expend energy for little to no return, but subsequently time spent
147 hunting will increase per consumed kill which can have detrimental impacts on overall
148 individual fitness [46]. Such apex predator effects can be so extreme that some mesopredators
149 have developed flexible behavioural strategies when handling prey, as well as spatial and
150 temporal measures, in order to avoid confrontations and coexists with apex predators within
151 the landscape [47, 48].

152 A whole suite of other species completes the facultative scavenger group. This includes many
153 omnivorous species that are not strictly predators and are highly opportunistic, these species
154 include corvid spp., racoon and possum spp., wild pigs, and many species of reptile [49]. These
155 (mostly) non-predatory species can be common scavengers in systems not dominated by
156 competitively superior scavengers [50-52]. However, in the presence of such dominant
157 scavengers, the scavenging rates of these non-predatory subordinate scavengers are often
158 suppressed, and this has consequently caused scavenging by such species to be underestimated
159 and even unknown [32]. Additionally, carrion is not exclusively a food resource but also a more
160 practical focal point within ecosystems for scavengers that may exploit other carcass resources.

161 For instance, many passerine species utilise carcasses as a source of hair or feathers for nesting
162 material [53]. Furthermore, carcasses also attract a host of invertebrate scavenger species, upon
163 which vertebrates, and indeed other invertebrates, can subsequently predate without explicitly
164 utilising the carcass [53].

165 **Invertebrate scavengers**

166 Invertebrate scavenger species are also a crucial component of any scavenger guild. In many
167 cases, invertebrate scavenging is much more complex than that of vertebrates, largely due to
168 the complexity of chemical and visual cues utilised by invertebrates for carrion detection,
169 colonization, and succession [54]. Terrestrial invertebrate scavenging is primarily limited to
170 insects which in turn are dominated by *Diptera* (true flies) and *Coleoptera* (beetles) but also
171 include *Hymenoptera* (ants, bees, and wasps), and *Acari* (mites) [54]. Scavenging by such
172 species orders can be highly contrasting on multiple temporal scales [55]. This is especially
173 apparent when considering the seasons, with carrion biomass loss attributed to insect
174 scavenging generally highest during summer and lowest during winter [54]. In some
175 ecosystems, this seasonal difference can be so extreme that during summer insect scavenging
176 is the primary driver of carrion decomposition rates, not vertebrate scavenging, despite the vast
177 amounts of carrion that vertebrates can consume at once [27, 56].

178 **Seasonality in scavenging ecology**

179 As previously noted, carrion is temporally patchy, and this is largely due to seasonality in the
180 carrion pool which is linked to animal deaths and predation rates [3, 7, 8]. Consequently,
181 scavenging is highly seasonal. The seasonal scavenging rates and activities of invertebrates can
182 largely be attributed to warm (high scavenging rates) and cold (low scavenging rates)
183 conditions [54]. However, the scavenging rates of vertebrates are more complex within this
184 seasonal framework due to contrasting seasonality in the available carrion pool and the

185 scavenging rates of the scavengers themselves [41]. For instance, it is generally accepted that
186 vertebrate scavenging rates are highest during winter, especially in the higher and lower
187 latitudes. This is due to a lack of alternative food sources and potentially more carcasses within
188 the landscape as many individuals succumb to the harsh conditions of winter [33, 42, 57-59].
189 However, it is also likely that during winter carcasses are harder to detect as olfactory cues
190 related to decomposition are lower due to reduced temperatures and decreased
191 invertebrate/microbial scavenging activity [51]. Further still, scavenging rates, activities, and
192 behaviours can also be linked to the life histories of vertebrate scavengers which are also often
193 linked to seasonal considerations such as breeding [60].

194 Seasonal effects (e.g., seasonality in the carrion pool, invertebrate scavenging, and vertebrate
195 scavenging) can have cascading impacts on scavenging dynamics that ultimately determine
196 how long carrion persist within ecosystems [61]. However, despite the strong consensus that
197 scavenging ecology is seasonal, many field-based studies still fail to sufficiently consider the
198 seasons, or replicate studies across all seasons [62]. A common field approach in scavenging
199 ecology is to sample and/or monitor only during two seasons of interest (e.g., hot and cold, or
200 wet and dry, or breeding and non-breeding) [32, 38, 63, 64]. This method overlooks the
201 importance of each of the seasons and simplifies the complexities of each, especially with
202 regards to scavenging dynamics that may occur in response to breeding, migrations, and/or
203 rapid vegetative change [32].

204 In response to this and other poor study designs, Schoenly et al. (2015) defined the successful
205 design of any robust field study in carrion ecology as those that simultaneously accounted for
206 temporal aspects, spatial aspects, and sample size, in conjunction with a suite of other minor
207 considerations [62]. Temporal considerations are important to account for seasonal effects and
208 variability over time [32, 62]. Spatial aspects can be equally as important when designing a
209 field study so that enough natural and environmental variation is covered in order to sufficiently

210 characterise a given ecosystem, as well as to consider spatial independence between monitored
211 carcasses. Sample size is especially important in scavenging ecology to ensure there is enough
212 independent units (i.e., carcasses) within a study to facilitate adequate seasonal replications and
213 ultimately yield statistically and ecologically valid and meaningful results.

214 **Scavenging in Australia**

215 Australia is home to a plethora of unique ecosystems that support equally unique scavenger
216 guilds. These diverse ecosystems exemplify the need to conduct ecological research in a
217 systematic manner for each differing ecosystem and across ecologically relevant temporal
218 (seasonal) and spatial scales [65]. This is particularly true given Australia's position in the
219 southern hemisphere and the literary bias for ecological (including scavenging) research in
220 northern hemisphere systems [66]. Further still, when considering the already highly variable
221 spatial, temporal, and interspecific nature of scavenging ecology, the Australian context only
222 exacerbates these complexities.

223 Despite this clear need for comprehensive studies, very little work has been done to extensively
224 describe scavenging ecology in Australia and understand its current context. Indeed, research,
225 thus far, has mostly focussed on imbedded topics within the scavenging field given the varied
226 motivations of different researchers and research groups. One area that has received
227 considerable attention is scavenging dynamics at the interface between the marine and
228 terrestrial environments. This is logical given Australia's extensive coastlines and potential for
229 harbouring considerable quantities of wave-swept carrion. Much of the work done here has
230 focused on the community structure and assemblage of beach scavenger guilds including
231 mammals, birds, and crustations, as well as how invasive species may impact the scavenging
232 dynamics of such guilds [67-73].

233 The impacts of invasive species have warranted widespread attention not only on Australia's
234 beaches and/or in the context of scavenging ecology, but also in many fields of ecology. In
235 addition to the beach-based studies mentioned above, extensive work has also been done on
236 invasive species including the cane toad (*Rhinella marina*), red fox (*Vulpes vulpes*), and feral
237 cat (*Felis catus*). Cane toads have invaded much of northern eastern Australia and are rapidly
238 expanding their range west and south [74, 75]. The invasive toads are highly toxic to native
239 predators and/or scavengers that consume them. Consequently, this toxicity has significant
240 potential to alter scavenging rates and scavenger guilds via a somewhat novel pathway that
241 considers carcass species rather than scavenger species [76, 77]. Red fox and feral cat
242 scavenging have also received some attention within Australian ecosystems for their abilities
243 to alter interspecific interactions and scavenging dynamics surrounding carrion [64, 67, 68,
244 78].

245 Aside from the impacts imposed by invasive scavengers, native apex predators are the other
246 highly influential scavenging taxa in Australia. Because of their status as a keystone species,
247 the dingo (*Canis dingo*) and Tasmanian devil (*Sarcophilus harrisi*) are dominant scavengers
248 on mainland Australia and Tasmania (an island state of Australia), respectively. Much of the
249 scavenging work on these species, with some exceptions [39, 64, 79, 80], has largely been
250 observational in nature, especially so for the dingo which, until recently, had been subject to
251 very few studies describing their scavenging activity [79, 81-83]. The Tasmanian devil is
252 recognised as a prolific scavenger, being one of the few terrestrial vertebrates globally (other
253 than vultures) thought to rely largely on a diet of carrion [84]. As such, the scavenger guilds of
254 Tasmania are highly unique, and their research has largely focussed on how the Tasmanian
255 devil drives scavenging dynamics [39, 84, 85]. Due to the transmissible devil facial tumour
256 disease (DFTD), and the associated catastrophic population declines, much of the work done
257 on the Tasmanian devil has focused on understanding the diseases properties. Consequently,

258 the scavenging rates and activities of Tasmanian devils remains relatively speculative [84, 86-
259 88], and indeed, existing scavenging based research has largely considered how such
260 population declines may change Tasmanian scavenger guilds and scavenging dynamics [39,
261 85].

262 Other areas covered with regards to scavenging in Australia include forensically important
263 scavenging activities [89, 90], lead poisoning of scavengers [91-94], anthropogenic carrion
264 subsidies [64, 78, 95, 96], and invertebrate scavenging dynamics [97-101]. With scavenging
265 ecology being a new and emerging field, the number of studies done, thus far, in Australia is
266 small. In addition, very few studies have been undertaken with adequate sample sizes and/or
267 accounted for spatial and temporal aspects [62]. It is also difficult to make inferences from the
268 findings of existing studies in the absence of the relevant natural history of scavenging in
269 Australia broadly, and more specifically for each of its ecosystems [102].

270 **Scavenging in the Australian Alps**

271 Despite only encompassing 0.16% of Australia's total land mass, the Australian Alps is a
272 unique and diverse ecosystem. This region has highly variable terrain ranging from steep
273 mountain peaks to flat alpine plains and includes many of the highest peaks and ranges within
274 Australia's Great Dividing Range, including Australia's highest peak – Mount Kosciuszko. As
275 a consequence of the terrain, the landscape has evolved to become delineated into three
276 altitudinally distinct ecological communities – montane (500m-1500m), subalpine (1500m-
277 1850m), and alpine (1850m-2228m) [66]. Each of these communities are highly seasonal with
278 the temperatures experienced throughout the course of a year differing by up to 50°C (-10°C –
279 40°C). During winter much of the landscape can be covered in snow, but by the following
280 summer the same landscape can experience intense bushfires, such as those during the 2019-
281 2020 Australian bushfire season [103]. With extreme weather events expected to become more

282 frequent and severe, many of Australia's native alpine flora and fauna species are at risk to a
283 rapidly changing climate in an already highly variable and vulnerable environment [104].

284 Despite being a highly volatile landscape, a vertebrate scavenger guild has become established
285 within the Australia Alps. This native guild is primarily composed of dingoes (*Canis dingo*)
286 and wedge-tailed eagles (*Aquila audax*) – apex predators; spotted-tail quolls (*Dasyurus*
287 *maculatus*) – mesopredator; brushtail possums (*Trichosurus vulpecula*), raven spp. (*Corvus*
288 *coronoides* and *Corvus mellori*), and pied currawongs (*Strepera graculina*) – facultative
289 scavengers. Whilst no study to date has holistically described the scavenging dynamics of the
290 Australian alpine guild, species specific scavenging has been described both within the Alps
291 and elsewhere in Australia [64, 79, 99, 105]. Invasive species also play a significant role within
292 the Australian alpine scavenger guild. Feral pigs (*Sus scrofa*), whose detrimental impacts are
293 largely attributed to ground rooting [106], are also pervasive scavengers in Australia's alpine
294 regions [99]. The invasive red fox and feral cat are likely two of the greatest threats to
295 Australia's native mammals, marsupials, and birds [107]. Both these invasive species,
296 however, do not exclusively hunt and are capable scavengers [64, 108]. In conjunction, the
297 detrimental impacts of these three invasive scavengers, could have widescale effects on
298 scavenging dynamics and the availability and persistence of carcasses within the Australian
299 Alps, especially so regarding feral pigs which are capable of consuming an entire carcass in
300 one scavenging bout. Many studies have demonstrated that dingoes can benefit the broader
301 Australian environment by suppressing red fox and feral cat populations despite their invasive
302 nature [109-111]. Therefore, a similar dynamic might be occurring in the Australia Alps,
303 specifically surrounding carrion. This could provide novel evidence for the regulation of an
304 invasive mesopredator population by a native apex predator within a scavenging context.

305

306

307 **Aims and hypotheses**

308 Carrion, in of itself, adds substance to the phrase ‘life after death’. This is because even after
309 the death of an animal its carcass can become a focal point within an ecosystem for a whole
310 host of organisms. It is for these reasons that carrion can be used to specifically examine
311 scavenging ecology. Within this thesis, the complex environmental dimensions (seasonality
312 and altitude) of the Australian Alps were utilised to conduct a comprehensive and systematic
313 analysis of this model ecosystems scavenging dynamics. To conform with the design principles
314 set out by Schoenly et al. (2015), a long transect was utilised to adequately capture enough
315 natural and altitudinal variability in the Australian Alps (spatial aspect). Along this transect, 15
316 spatially independent carcasses were monitored per season (60 total), to account for seasonal
317 effects (temporal aspect), and this yielded statistically and ecologically relevant findings of
318 scavenging dynamics. Therefore, this thesis provides one of the first comprehensive analyses
319 of scavenging dynamics for an understudied Australian ecosystem and demonstrates a robust
320 and repeatable study design valuable to understanding scavenging ecology globally.

321 The second chapter of the thesis will investigate how the seasons affect scavenging dynamics
322 in the Australian Alps. Specifically, seasonal effects will be assessed to determine how they
323 affect four vertebrate scavenging variables: scavenger species richness/composition, time to
324 first arrival and scavenging at a carcass, scavenger activity (i.e., probability of a scavenger
325 investigating vs scavenging a carcass), and time spent investigating and scavenging a carcass.
326 Broadly, it is predicted that the seasonal effects will significantly impact most if not all of the
327 scavenging variables, most notably during summer and winter when thermal extremes are at
328 their peak. More detailed hypotheses are provided in *Chapter 2*.

329 The third chapter explores the top-down scavenging effects of an apex predator. Much of the
330 literature describes the interspecific interactions of apex predators with mesopredators [25, 40,
331 45, 46, 48, 112, 113]. However, little is known about the interspecific interactions apex

332 predators have with subordinate facultative scavengers considered potential prey species. In
333 the Australian Alps, the dingo is the apex (scavenging) predator, and thus, may be exerting top-
334 down scavenging effects on subordinate facultative scavengers such as the brushtail possum
335 and raven spp., which were demonstrated to be the most common scavengers in *Chapter 2*. It
336 is expected that brushtail possums, and to a lesser degree raven spp., may use a suite of spatial,
337 temporal, and behavioural methods to avoid confrontations with dingoes surrounding carrion.
338 More detailed hypotheses are provided in *Chapter 3*.

339 **Implications**

340 In addition to providing a comprehensive account for the scavenging dynamics of the
341 Australian Alps, the findings of this thesis are relevant to environmental managers. The
342 Australian Alps are home to a plethora of invasive animal species including feral
343 horses/brumbies (*Equus ferus caballus*), various deer species, feral pigs (*Sus scrofa*), red foxes
344 (*Vulpes vulpes*), feral cats (*Felis catus*), feral goats (*Capra hircus*), and feral rabbits
345 (*Oryctolagus cuniculus*). As such, extensive management practices are routinely undertaken to
346 control the populations of these species and their negative impacts on the fragile Australian
347 alpine ecosystem. Often, many of these operations leave culled animal carcasses *in situ* to
348 decay, and this provides a sudden influx of available energy and nutrient within the landscape.
349 Therefore, the Australian Alps may at times harbour high rates of carrion biomass.

350 Scavengers provide vital ecosystem services valuable to both the natural and human
351 environment, the latter of which is often overlooked. To take advantage of these services, the
352 findings of this thesis may assist in determining when (i.e., which season) invasive species
353 management practices should occur. Firstly, many of the key invasive species in the Australian
354 Alps (feral pigs, red foxes, and feral cats) are scavengers. Therefore, management programs
355 have the potential to supplement and support such invasive species populations by increasing
356 carrion biomass, and thus, nullifying control efforts. Using my findings to determine the

357 seasons in which native scavenging rates and invasive scavenging rates are highest and lowest
358 respectively may reduce the chance of such a counterintuitive result. Secondly, in addition to
359 the natural/environmental features of the Australian Alps there is also a complex human
360 dimension. The Australian Alps is home to many popular recreational pursuits and as such has
361 become a tourism hotspot within Australia [114]. Therefore, the faster processing and
362 decomposition times of carcasses, as facilitated by the scavenger guild during their most
363 efficient seasons, would be beneficial to the overall environmental aesthetics of the Australian
364 Alps as perceived by humans.

365 This thesis will provide an ecological context for the natural history of scavenging dynamics
366 in the Australian Alps. Using this ecosystem, one previously ignored, our findings will
367 supplement the relatively new literary foundation of scavenging ecology. In addition to the
368 ecological findings, it is also anticipated that this thesis will set a new standard for field-based
369 scavenging research. The methods used here, specifically regarding the temporal scales and
370 sample sizes used, if adopted by future studies at a minimum, can adequately describe localised
371 scavenging dynamics within any given ecosystem, thus, facilitating the advancement and our
372 understanding of scavenging ecology globally.

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656

1 **Chapter 2 – Carcass use by mesoscavengers drives**

2 **seasonal shifts in scavenging dynamics**

3 **Abstract**

4 Carrion is a high energy and nutrient rich resource that attracts a diverse group of vertebrate
5 scavengers and consequently can be utilised as a focal point within an ecosystem to assess
6 scavenging dynamics. Despite the carrion pool being highly seasonal, many studies utilising
7 carrion to investigate scavenging dynamics, have neglected to account for seasonal effects.
8 Therefore, using the highly seasonal Australian alpine ecosystem, carrion deployed
9 experimentally each season were utilised to assess vertebrate scavenging dynamics. Time to
10 first detection of carcasses by vertebrate scavengers was longer during summer, likely a
11 product of increased invertebrate scavenging rates and an abundance of other available food
12 sources. Scavenging was more likely than investigation of carcasses during winter and spring.
13 During winter, this trend was driven by an increase in brushtail possum scavenging which
14 accounted for 78% of all scavenging events and was likely in response to a scarcity of other
15 food sources. Conversely, during spring, scavenging was more likely as the scavenging rates
16 of raven species increased four-fold to meet the demands of breeding (i.e., increased energetic
17 requirements, nest construction, and chick rearing). The frequent but brief trips between nest
18 and carcass by raven spp. during the breeding seasons was also likely the reason scavenging
19 events were significantly shorter during spring. These results demonstrate highly seasonal
20 scavenging dynamics in the Australian Alps and supports the importance of seasonal effects in
21 scavenging ecology. However, the findings are unique given seasonal trends in scavenging
22 dynamics were dominated by low-ranking mesoscavengers: the brushtail possum and raven
23 species.

24 **Introduction**

25 Carrion is a valuable ecosystem resource which in contrast to plant-based detritus is high in
26 energy and nutrient rich [1]. Although carrion is spatially and temporally patchy [2, 3], it is
27 exploited by species that have evolved to scavenge. Historically, scavengers have been viewed
28 as ‘bottom-feeders’, due to associations with rotting matter, disease, and death [4]. But
29 scavenging is present in most taxa, including obligate and facultative large vertebrate
30 scavengers capable of consuming a whole carcass in one feeding event, to invertebrate
31 scavengers that can aggregate around carrion in the thousands [5]. Together, these species form
32 scavenger guilds, which in addition to acting as ‘natures clean-up crew’ [6], are also important
33 to supporting critical linkages, structure, and stability in food webs [2]; distributing nutrients
34 within and among ecosystems; and providing economic and human health benefits related to
35 carcass disposal and sanitary measures [7].

36 Carrion biomass within an ecosystem fluctuates in response to key modes of death such as
37 predation, but also in response to environmental factors [5, 8]. Seasons are one of the strongest
38 governing environmental forces and can dictate the life histories of many herbivorous and/or
39 migratory species that form a major component of the carrion pool [5, 8-10]. This is because,
40 in highly seasonal environments carrion biomass can become cyclically pulsed towards the
41 ends of harsh and/or prolonged seasons when such species incur increased mortality rates [11].
42 This is evident in tropical and sub-tropical Africa following wet and dry season cycles [12-14].
43 For example, each year over 1 million wildebeest (*Connochaetes taurins*) migrate through the
44 Serengeti Mara ecosystem [15], including through the Mara River, where mass wildebeest
45 drownings occur during its crossing [16, 17]. This sudden annual influx of carrion provides
46 many terrestrial and aquatic scavengers with an abundance of available energy and nutrients
47 [18]. Similar trends are also common in the higher northern latitudes where many ungulate
48 species face increased mortality rates towards the ends of harsh northern winters [19-21].

49 Season, along with daily temperatures, humidity, and moisture levels can also influence carrion
50 persistence rates via the effects they have on regulating microbial and insect activity [22-24].

51 It is in response to such seasonal fluctuations in the availability of carrion, that scavengers have
52 adapted and evolved to exploit such a pulsed food resource. Indeed, facultative scavengers have
53 considerable flexibility in their diets regarding both the relative contribution of predated vs
54 scavenged food and meat vs other food sources [25]. For example, the Białowież Forest
55 (Poland) scavenger guild demonstrated increased scavenging rates during winter in response
56 to greater energy expenditure (i.e., keeping warm, traversing snow), carcass availability, and
57 lack of other food sources [26-28]. These dynamics can become even more complex when
58 considering the contrasting seasonality in the life histories of each scavenger species within a
59 guild. For instance, during peak chick rearing season, the energy requirements of herring gulls
60 (*Larus argentatus*) and lesser black-backed gulls (*Larus fuscus*) are inherently greater, and
61 consequently both species were less selective of fish species when scavenging fishery discards
62 [29]. A plethora of other environmental, life history, and inter/intra specific factors can also
63 affect species specific scavenging, and thus, add further complexities to scavenging dynamics
64 [30].

65 Despite obvious seasonal trends in scavenging ecology, the experimental designs of many
66 field-based scavenging studies often overlook the impact of seasonal effects [31]. Indeed, it is
67 common to monitor scavenging dynamics surrounding carrion only during two seasons of
68 interest (i.e., hot and cold, or wet and dry, or breeding and non-breeding) [23, 32-36]. Such an
69 approach potentially oversimplifies the ways in which the various dimensions of scavenging
70 ecology can change not only between each of the seasons, but over the course of a year, and
71 even between years [36]. Moreover, some studies completely ignore seasonal effects by only
72 monitoring scavenging dynamics during one season [37-41]. The successful design of any
73 robust field study in carrion ecology should simultaneously accounts for temporal aspects (i.e.,

74 diel, seasonal, yearly), spatial aspects (i.e., representative of ecosystem, spatially independent),
75 and sample size, in conjunction with a suite of other minor considerations [31].

76 In this study, the highly seasonal nature of the Australian Alps was exploited to monitor the
77 use of carcasses by vertebrate scavengers. Carcass monitoring was replicated across all four
78 seasons to account for seasonal effects. Such an approach, whilst accounting for environmental
79 variability, provided an opportunity to determining how important the seasons are in
80 influencing scavenger species richness/composition, and to test specific predictions related to
81 (1) the time to first detection and scavenging at a carcass; (2) scavenger activity (i.e.,
82 probability of a scavenger investigating vs scavenging a carcass); and (3) the time spent
83 investigating and scavenging a carcass. Accelerated decomposition of carrion during summer
84 as a product of increased invertebrate and microbial activity is known to produce stronger
85 carcass-linked odours [42]. Consequently, it was predicted that increased olfactory cues during
86 summer would result in greater detectability of carrion by vertebrate scavengers, and thus,
87 shorter time to first arrival and scavenging at carcasses. Conversely, despite carrion being less
88 detectible during winter, it was predicted that the probability of vertebrate scavenging would
89 increase, and time spent scavenging would be longest. This prediction was informed by the
90 findings of previous scavenging studies that demonstrate vertebrate scavengers to rely on
91 carrion more heavily during winter when other food resources are scarce [26-28]. The results
92 are used to highlight the importance of replicating field-based scavenging studies across the
93 seasons in order to fully understand the complex scavenging dynamics and interactions that
94 take place surrounding carrion.

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98 **Methods**

99 **Study site**

100 This study was conducted in Kosciuszko National Park, located in southern New South Wales,
101 Australia. This region includes many of the highest peaks and ranges within Australia's Great
102 Dividing Range, including Australia's highest peak – Mount Kosciuszko. The landscape is
103 delineated into three altitudinally distinct ecological communities – montane (500m-1500m)
104 subalpine (1500m-1850m), and alpine (1850m-2228m) [43]. This work was undertaken within
105 the montane zone (between approximately 1000m – 1500m) which is characterised by forest
106 stands dominated by snow gum (*Eucalyptus pauciflora*) in association with various other
107 *Eucalyptus* species.

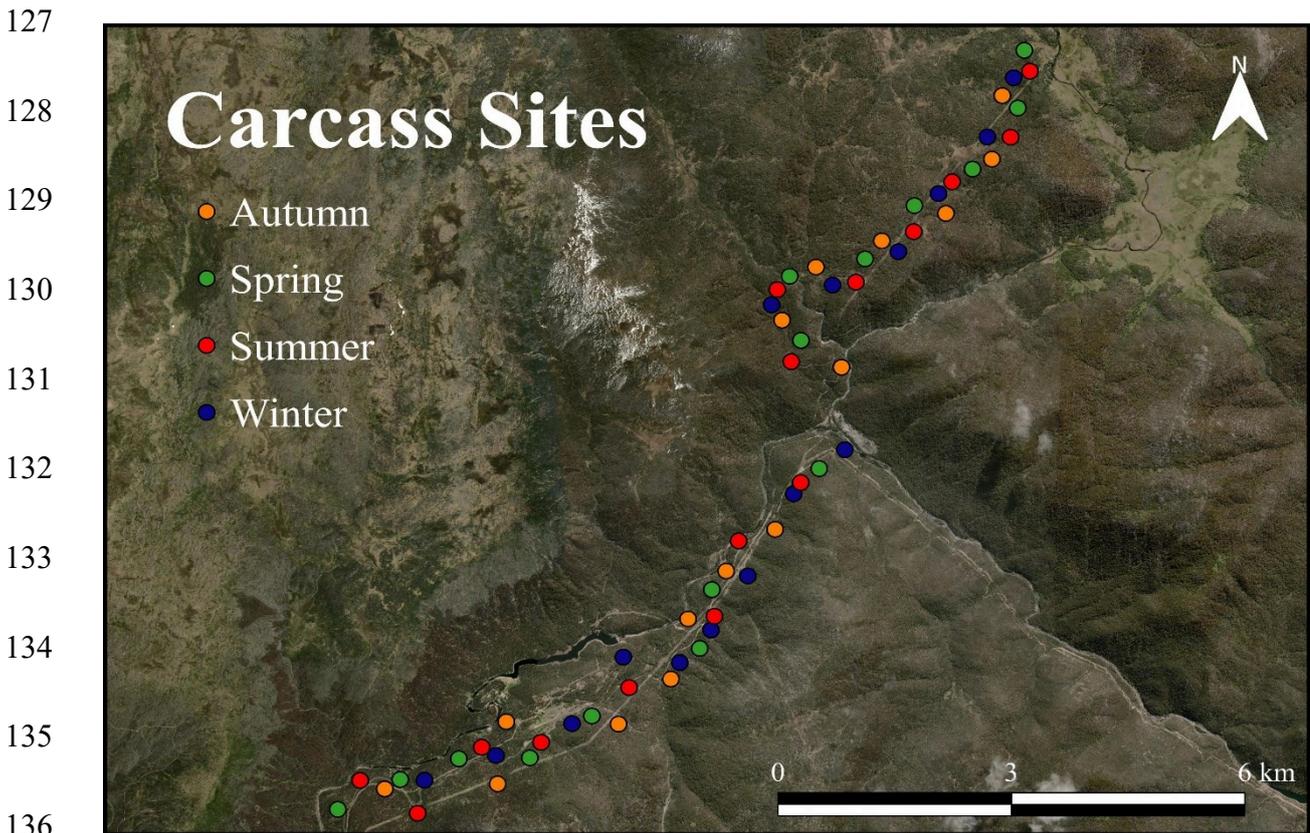
108 **Ethics, licenses, and permits**

109 The following described work received all required ethics, licenses, and permits approved by
110 the relevant authorities (i.e., The University of Sydney; New South Wales Office of
111 Environment and Heritage; and New South Wales National Parks and Wildlife Services). All
112 kangaroo carcasses used for the purposes of this research were sourced fresh and locally from
113 existing authorised and legally approved management culls that are conducted to control
114 overabundant kangaroo populations.

115 **Fieldwork**

116 A 15 km transect was established through Kosciuszko National Park along which all carcass
117 monitoring took place (Figure 1). This transect ran northeast – southwest from a border region
118 of the national park inwards towards its interior and was selected due to its accessibility (i.e.,
119 road access) and because it is a relatively undisturbed area with little human activity.
120 Monitoring periods were established to coincide with the four seasons: autumn – March 2020;
121 winter – July 2020; spring – October 2020; summer – January 2021. During each season, 15

122 sites were established along the transect, separated by approximately 1 km from the nearest
123 sites monitored within the same season and approximately 250m from the nearest sites
124 monitored during other seasons (60 different sites in total; Figure 1). The separation of sites
125 ensured a level of spatial independence and it prevented habituation of scavengers to a carrion
126 source location [35, 44-46].



137 **Figure 1.** Location of the monitoring transect along which all 60 carcass sites were established
138 within Kosciuszko National Park (red = summer, orange = autumn, blue = winter, green =
139 spring).

140 Within each seasonal monitoring period, one fresh eastern grey kangaroo carcass was placed
141 at each site (15 total). The carcasses ranged in weight from 10kg-70kg and on average were
142 28.3kg (± 1.498 – standard deviation). Each carcass was secured, using wire ties, to star pickets
143 driven into the ground to ensure they remained in situ to be monitored for 60 days [44-46].
144 Vertebrate scavenger activity was monitored at each site using a Reconyx PC800 Hyperfire™

145 remote camera. Each camera was placed on a free-standing star picket three meters north of
146 the carcass – the southern aspect of the cameras decreased exposure to direct sunlight which
147 would otherwise reduce image quality. The cameras were calibrated to take photographs
148 continuously (approximately one image per second) when triggered by thermal movement (i.e.,
149 rapid-fire, no wait period). These approaches and methods follow those previously used in
150 field-based scavenging research [21, 34, 35, 44-48].

151 **Analyses**

152 Remote camera images were analysed for species presence and the number of individuals of a
153 species present. In order to determine distinct visitations of species, an ‘event’ was
154 characterised as a visitation by a species that occurred more than 10 minutes after the last
155 visitation by that same species. Only species-specific events could be characterised because
156 identification of individuals for most species was not possible. An event was characterised as
157 a ‘scavenging event’ if the species present scavenged on the monitored carcass in at least one
158 of the remote camera images consisting of that event, otherwise the event was characterised as
159 an ‘investigation event’. Data from species recorded to have scavenged at least once were
160 included in the statistical analyses.

161 The R software environment (version 1.4.1717) was used for all statistical analyses. Moran’s I
162 statistic was utilised to test for spatial autocorrelation in each of the scavenging response
163 variables between the carcass sites (R Package ‘ape’; [49]). To determine any differences in
164 species composition between the seasons, a permutational multivariate analysis of variance
165 (PERMANOVA; R Package ‘vegan’; [50]) was used in conjunction [51] with an analysis of
166 similarities (ANOSIM; R Package ‘vegan’; [50]). To determine which scavenger species were
167 driving any differences in species diversity between the seasons a similarity percentages
168 (SIMPER) analysis was used (R Package ‘vegan’; [39]).

169 To adequately characterise the seasonal nature of scavenging, four response variables were
170 used: scavenger species richness (Conway-Maxwell Poisson distribution), the time to first
171 detection of and scavenging at a carcass (in hours; Gamma distribution), scavenger activity
172 (i.e., probability of a scavenger investigating vs scavenging a carcass; binomial distribution),
173 and investigation and scavenging event duration (in minutes; Gamma distribution). Each of the
174 response variables were modelled against the explanatory variables of season, as well as
175 altitude to account for any differences in elevation, however, given the small altitudinal
176 gradient used (500m) no altitudinal effects were expected. These models were constructed
177 twice, once using only investigation events and once using only scavenging events. Only one
178 model was constructed for scavenger activity as it is a binomial response variable (either
179 investigation; 0, or scavenging; 1) designed specifically to determine the probability of either
180 an investigation event or scavenging event.

181 The relationships of each of these scavenging response variables with the explanatory variables
182 of season and altitude were modelled using either generalised linear models (GLM; R Package
183 ‘lme4’ [52]) or generalised linear mixed models (GLMM; R Package ‘lme4’ [52]), and in the
184 case of poorly fit models, generalised additive models were utilised (GAM; R Package ‘mgcv’
185 [53]). To determine the most parsimonious model(s), Akaike information criterion (AIC) [54]
186 was used ($\Delta AICc$ level of significance < 2), with model selection facilitated by the utilisation
187 of the dredge function (R Package ‘MuMIn’ [55]). AIC considers the different combinations
188 of explanatory variables (i.e., combinations of season and altitude) within a model and as such
189 the scavenging response variables had four possible models: non-interaction season and
190 altitude model ($x \sim y + z$), season model ($x \sim y$), altitude model ($x \sim z$), and null model ($x \sim 1$).
191 Significance testing (p level of significance < 0.05) was also undertaken using the base model
192 (i.e., non-interaction season and altitude model – $x \sim y + z$) to determine which explanatory
193 variables (seasons and/or altitude) and/or their levels (summer, autumn, winter, spring) were

194 important in explaining each of the scavenging response variables. To yield additional
195 information from these models, Tukey's honest significance tests were used to determine which
196 seasons were significantly different from one another regarding the modelled scavenging
197 response variable (R Package emmeans) [56]. This pair-wise test approach could not be used
198 when modelling species richness because it followed a Conway-Maxwell-Poisson distribution
199 which does not support post-hoc analyses. However, the previously described PERMANOVA,
200 ANOSIM, and SIMPER analyses provided similar relevant insights.

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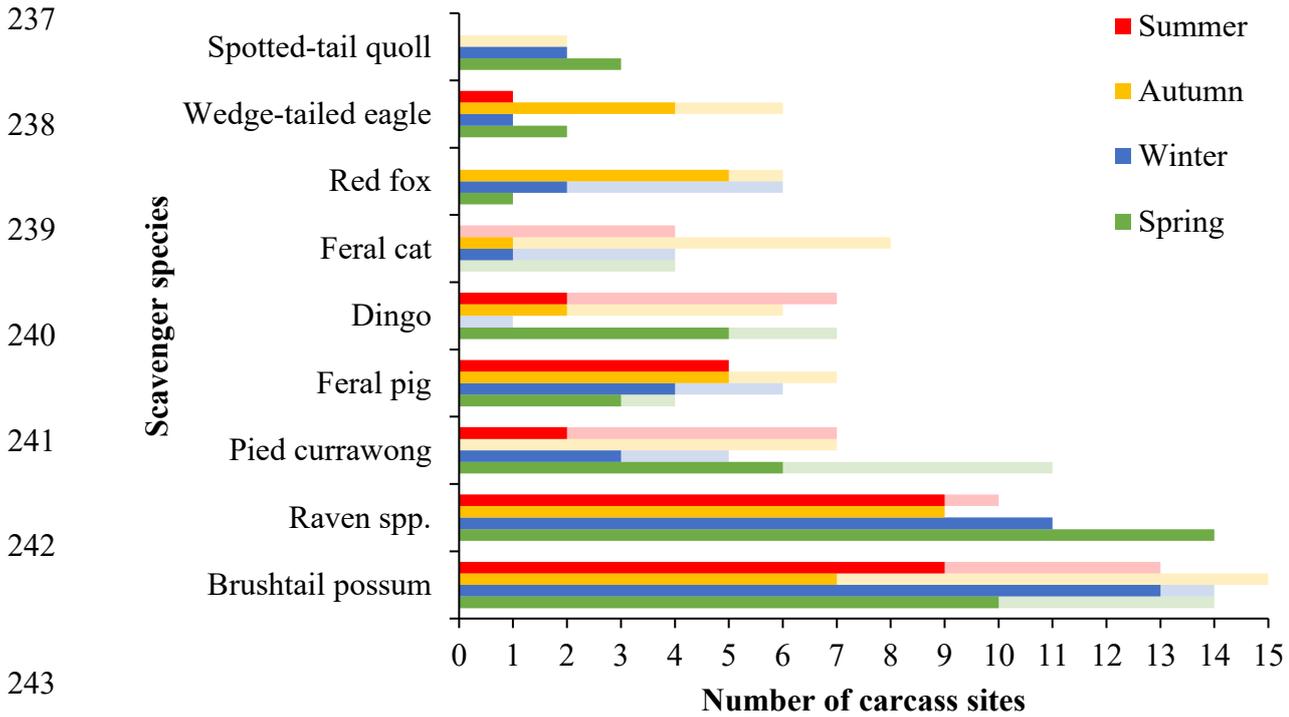
211 **Results**

212 Of the 60 carcass sites monitored, remote camera data was gathered for 58; remote camera data
213 for two sites (one during winter and one during summer) were lost due to theft and camera
214 failure. The camera traps took 745,599 images of 34 different species including both scavenger
215 and non-scavenger species. Of these species nine were considered scavenger species based on
216 recorded scavenging of the monitored carcasses (Figure 2). These were the: spotted-tail quoll
217 (*Dasyurus maculatus*), feral cat (*Felis catus*), dingo (*Canis dingo*), pied currawong (*Strepera*
218 *graculina*), wedge-tailed eagle (*Aquila audax*), brushtail possum (*Trichosurus vulpecula*),
219 raven spp. (*Corvus coronoides* and *Corvus mellori* – indistinguishable from one another in the
220 remote camera images), red fox (*Vulpes vulpes*), and feral pig (*Sus scrofa*).



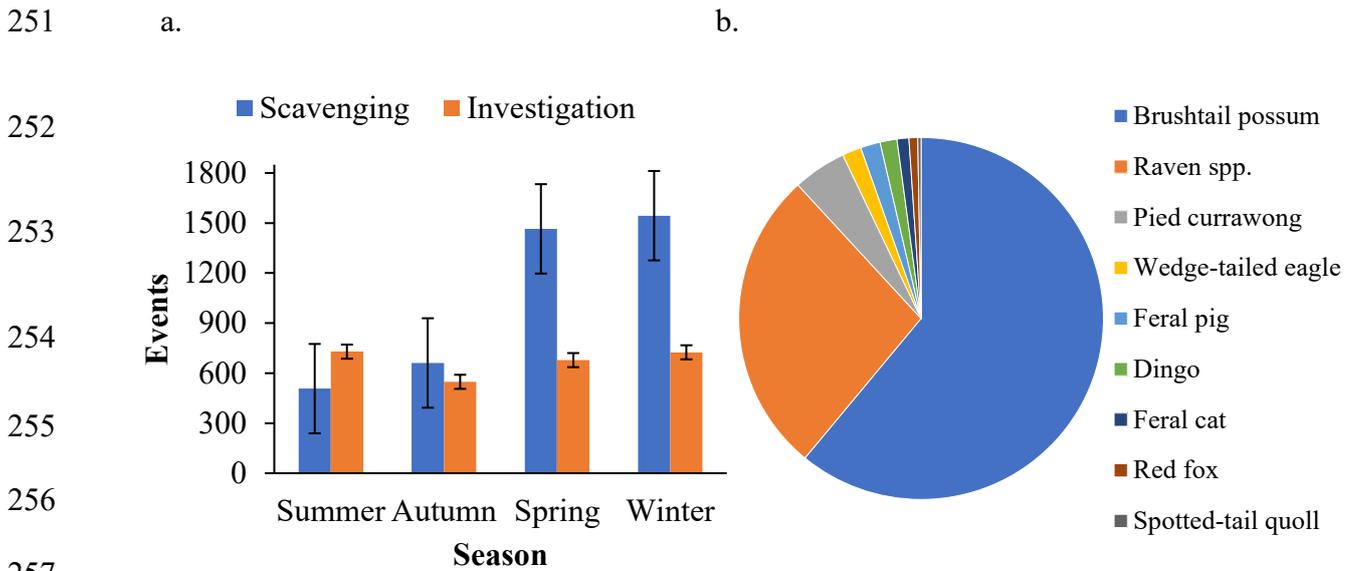
230 **Figure 2.** Remote camera images for each of the members of the Australian alpine scavenger
231 guild, determined by recorded scavenging of carcasses in Kosciuszko National Park.

232 These scavenger species occurred at carcass sites to investigate or scavenge carcasses at
 233 varying rates across the seasons (Figure 3). In total 6857 distinct events were recorded of which
 234 2680 were investigation events and 4177 scavenging events (Figure 4.a). Brushtail possums
 235 and raven spp. accounted for 88% of the total recorded events, whilst spotted-tail quolls
 236 accounted for the fewest events (Figure 4.b).



244 **Figure 3.** The total number of carcass sites visited by each scavenger species per season (red
 245 = summer, autumn = yellow, winter = blue, spring = green) where carcass sites that were
 246 recorded to have been scavenged (solid fill) are distinguished from those that were only
 247 investigated (shaded fill).

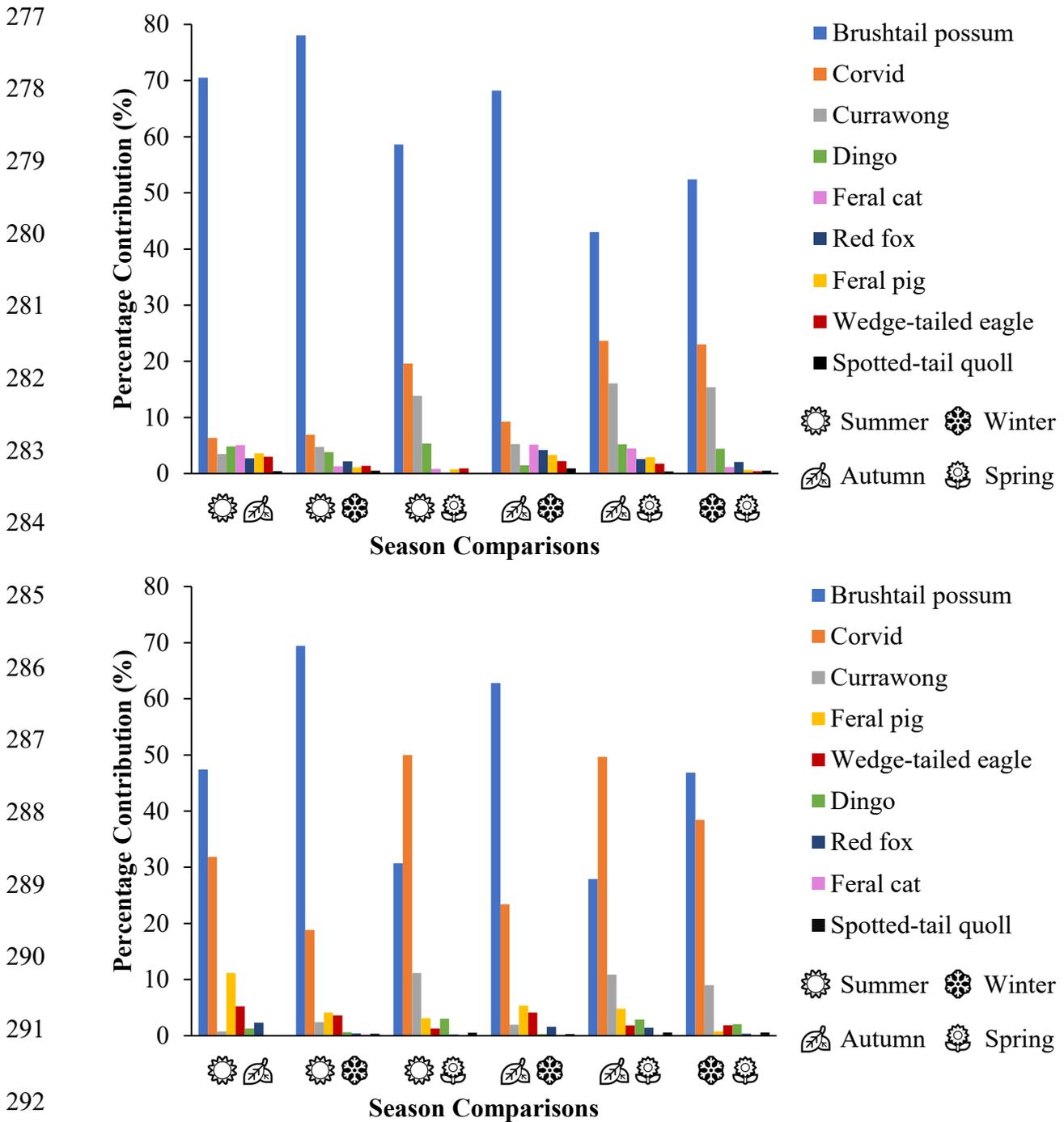
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258 **Figure 4. (a.)** The total number of events (orange = investigation, blue = scavenging) per
 259 season and **(b.)** the percentage of total events attributed to each scavenger species where
 260 brushtail possums accounted for 61%, raven spp. 27%, pied currawongs 5%, and all other
 261 species < 2% each.

262 No spatial autocorrelation was detected in any of the scavenging response variables (Table S1).
 263 The PERMANOVAs demonstrated that there were differences in species composition between
 264 the seasons for both investigation events ($p = 0.001$) and scavenging events ($p = 0.004$) but that
 265 they were weak differences ($R^2 = 0.146$ and 0.152 respectively; Table S2). The Tukey's honest
 266 significance tests for the investigation events PERMANOVA demonstrated that species
 267 composition was significantly different between autumn and spring ($p = 0.030$) and between
 268 spring and winter ($p = 0.018$; Table S3). The Tukey's honest significance tests for the
 269 scavenging events PERMANOVA demonstrated that species composition was significantly
 270 different between autumn and winter ($p = 0.036$) and between spring and winter ($p = 0.012$;
 271 Table S3). Furthermore, in conjunction with the PERMANOVA, the investigation events ($R =$
 272 0.109 with $p = 0.001$) and scavenging events ($R = 0.109$ with $p = 0.002$) ANOSIM analyses
 273 suggested that differences in species composition within the seasons and between the seasons

274 was more or less the same (Table S4). The SIMPER analysis demonstrated that brushtail
 275 possum and raven spp. were the primary drivers of the observed differences in species
 276 composition between the seasons (Figure 5 & Table S5/S6).



293 **Figure 5.** The percentage contribution of each scavenger species (contributed most to least, top
 294 to bottom of each respective legend) to the differences observed in species composition
 295 between each of the seasons for investigation events (top) and scavenging events (bottom). The

296 percentage contribution was adapted from the ‘cumulative sum’ results yielded from the
297 SIMPER analyses (Table S5 and S6).

298 **Scavenger species richness for investigation events**

299 All four models were parsimoniously competitive ($\Delta\text{AICc} < 2$) in explaining the variation in
300 scavenger species richness for investigation events (Table S7). The base model demonstrated
301 that summer ($p = 0.040$) and winter ($p = 0.035$) had a significant effect on scavenger species
302 richness for investigation events (Table S8).

303 **Scavenger species richness for scavenging events**

304 Three models were parsimoniously competitive ($\Delta\text{AICc} < 2$) in explaining the variation in
305 scavenger species richness for scavenging events (Table S7). The first was the null model
306 ($\Delta\text{AICc} = 0.00$), the second the seasons model ($\Delta\text{AICc} = 1.01$), and the third the altitude model
307 ($\Delta\text{AICc} = 1.26$). The base model was not significant (Table S8).

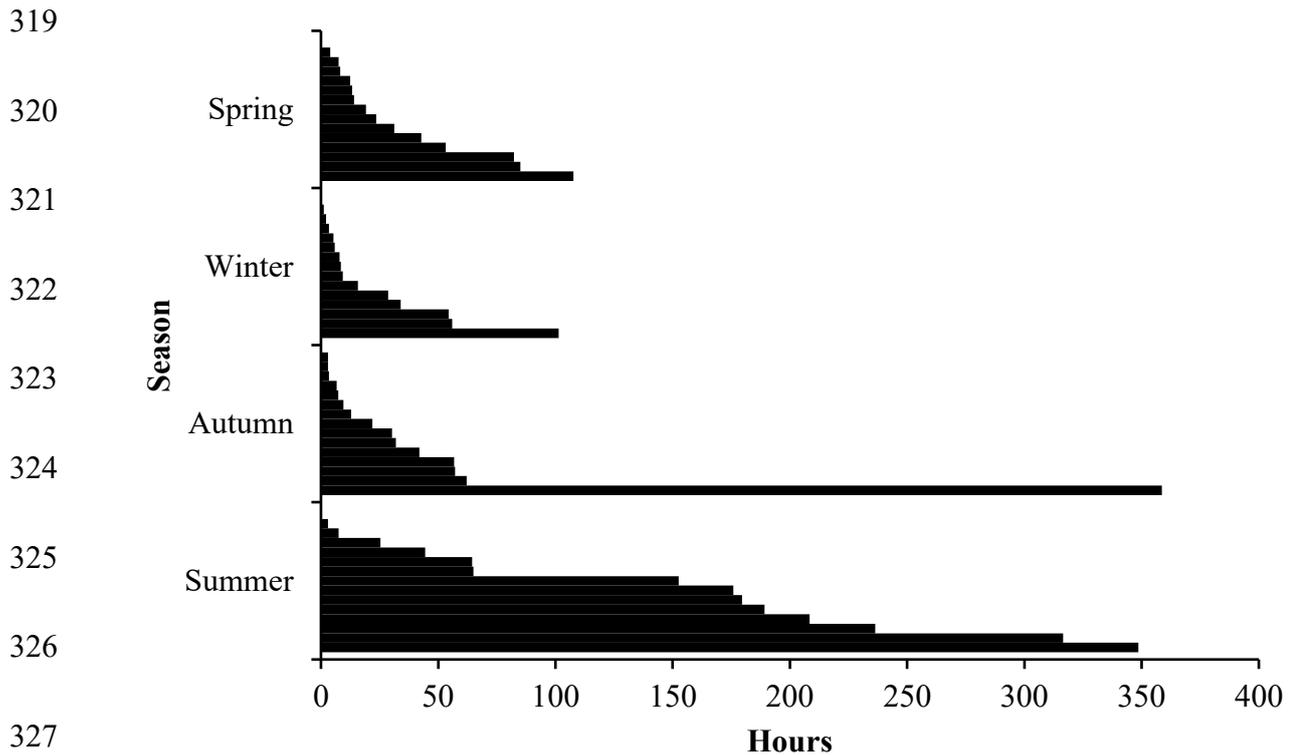
308 **Time to first detection of carcasses**

309 Two models were parsimoniously competitive ($\Delta\text{AICc} < 2$) in explaining the variation in time
310 to first detection of carcasses (Table S9). The first was the season model ($\Delta\text{AICc} = 0.00$) and
311 the second was the non-interaction season and altitude model ($\Delta\text{AICc} = 1.99$). The base model
312 demonstrated that summer had a significant effect on time to first detection of carcasses ($p =$
313 0.002 ; Figure 6; Table S10). Specifically, time to first detection of carcasses (in hours) was
314 4.289 and 6.527 times longer during summer than spring ($p = 0.012$) and winter ($p = 0.0010$)
315 respectively (Table S11).

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328 **Figure 6.** Hours to first detection of each carcass monitored during each of the seasons.

329 **Time to first scavenging of carcasses**

330 Three models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in time
 331 to first scavenging of carcasses (Table S9). The first was the season model ($\Delta AICc = 0.00$), the
 332 second the null model ($\Delta AICc = 0.65$), and the third the non-interaction season and altitude
 333 model ($\Delta AICc = 1.93$). The base model demonstrated that winter had a significant effect on
 334 time to first scavenging of carcasses ($p = 0.039$; Table S10).

335 **Scavenger activity**

336 Two models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in
 337 scavenger activity (Table S12). The first was the season model ($\Delta AICc = 0.00$) and the second
 338 was the non-interaction season and altitude model ($\Delta AICc = 1.61$). The base model
 339 demonstrated that winter ($p = 0.011$) and spring ($p = 0.003$) had a significant effect on
 340 scavenger activity (Figure 4; Table S13). Specifically, scavenging was 2.173 and 3.108 times

341 more likely than investigation during spring than autumn ($p = 0.015$) and summer ($p = < 0.001$)
342 respectively, and 2.787 times more likely during winter than summer ($p = < 0.001$; Table S14).

343 **Duration of investigation events**

344 Two models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in the
345 duration of investigation events (Table S15). The first was the null model ($\Delta AICc = 0.00$) and
346 the second was the season model ($\Delta AICc = 1.89$). The base model was not significant (Table
347 S16).

348 **Duration of scavenging events**

349 Two models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in the
350 duration of scavenging events (Table S15). The first was the season model ($\Delta AICc = 0.00$) and
351 the second was the non-interaction season and altitude model ($\Delta AICc = 0.86$). The base model
352 demonstrated that spring had a significant effect on the duration of scavenging events ($p = <$
353 0.001 ; Figure 7; Table S16). Specifically, scavenging event duration (in minutes) was 1.895
354 and 1.493 times shorter during spring than autumn ($p = < 0.001$) and winter ($p = 0.041$)
355 respectively (Table S17).

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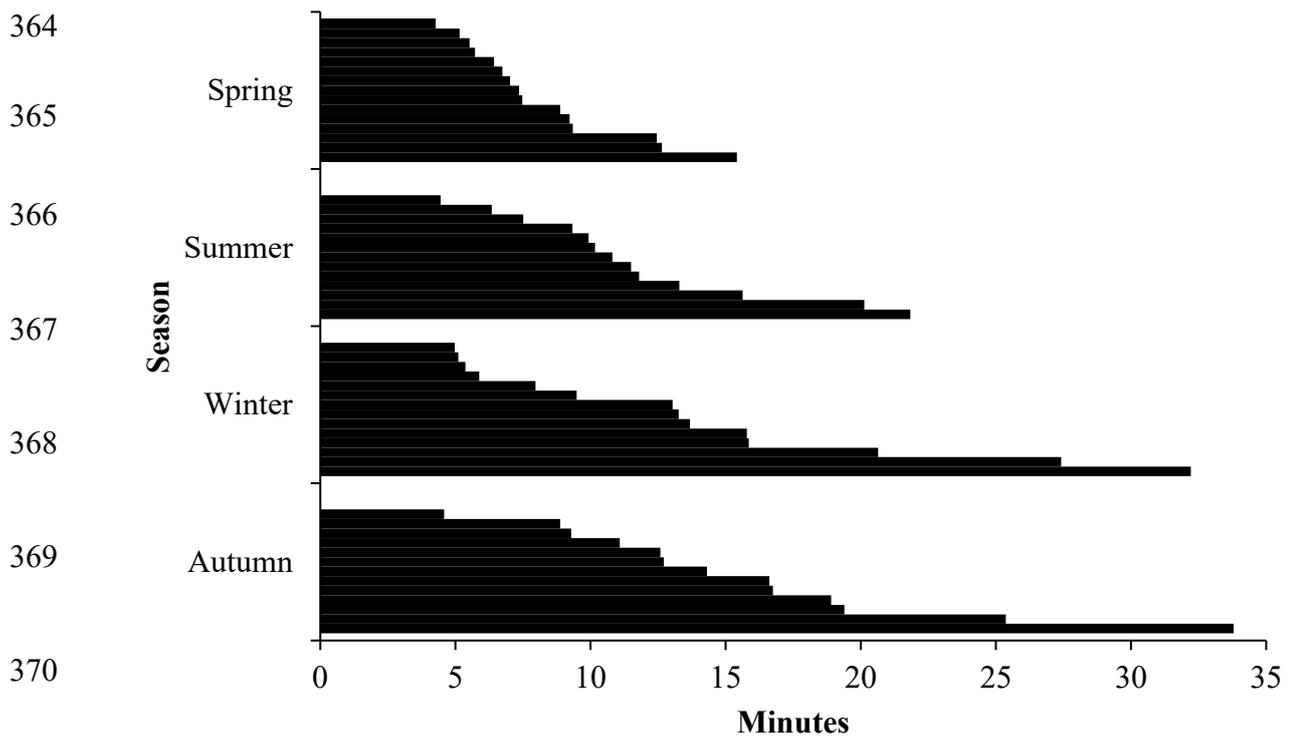
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371 **Figure 7.** The average duration of scavenging events (minutes) at each carcass monitored
 372 during each of the seasons.

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381 **Discussion**

382 This study determined how seasons affected vertebrate scavenging dynamics. It was expected
383 that scavenging dynamics would be highly seasonal, and thus, highlight the importance of
384 seasonal effects, a consideration often absent from the experimental designs of field-based
385 research in scavenging ecology. The findings here demonstrate that scavenging dynamics were
386 highly seasonal, and this trend was consistent (i.e., did not significantly change) across the
387 altitudinal gradient where the carcasses were monitored. Shifts in scavenging dynamics were
388 overwhelmingly dictated by the scavenging rates and activities of mesoscavengers, specifically
389 brushtail possums and raven spp.. The scavenging rates observed by raven spp. were not
390 unusual, with many studies both within Australia and globally describing raven spp. species as
391 prolific scavengers [27, 44, 45, 57, 58]. However, the brushtail possum has received little
392 recognition as a regular scavenger [59-61].

393 Australia has no true obligate vertebrate scavenger, but our study recorded carcass use by nine
394 species of facultative scavenging vertebrates. Scavenging rates by these species should be
395 highly seasonal, linked to factors such as carrion biomass, availability of other food sources,
396 and/or the life histories of the scavenger species [12-14, 20, 25-28, 62-64]. This was supported
397 by the results, but the trends uncovered were driven by the scavenging behaviours of the
398 brushtail possum and raven spp who together accounted for 88% of all recorded events.

399 Scavenging dynamics surrounding carrion are not typically dictated by the species-specific
400 scavenging rates of mesoscavengers, but that of larger dominant scavenger species [20, 65-67].
401 This is because larger species are generally more competitively dominant surrounding carrion
402 [11, 25, 35], and are able to open up carcasses, and thus, provide access to smaller scavengers
403 unable to do so [27, 68, 69]. In this study, however, larger scavengers such as dingoes, wedge-
404 tailed eagles, and feral pigs only accounted for 5% of all recorded events, suggesting they were

405 either not abundant during the study period or were not scavenging frequently. At carcass sites
406 they did visit, there was evidence that they could rapidly consume the carcass biomass,
407 including bones; in one case a dingo was observed consuming an entire kangaroo carcass
408 within a 24-hour period. Whether the relative absence of larger scavengers at carcass sites in
409 this study aided mesoscavenger access to the food resource is unknown but could reasonably
410 be expected if this resulted in less competition for the food resource and/or reduced predation
411 risk for the mesoscavengers. The relative absence of dingoes at carcass sites, along with
412 similarly low rates of scavenging by red foxes and feral cats, is likely to have especially
413 influenced the use of carcasses by brushtail possums, as they regularly feature in the diets of
414 these three predators [70-79].

415 The time it takes scavengers to detect a carcass, and subsequently scavenge it, is intrinsically
416 linked with carcass decomposition rates and persistence within the ecosystem [80]. In our
417 study, carcasses took longer to be detected during summer, especially when compared to spring
418 and winter. Carcass detection by scavengers is dependent on a number of factors including
419 olfactory cues, visual cues, inter/intra specific cues, and search effort [36, 42, 81]. We
420 therefore, expected that olfactory cues would be the primary mode of detection given that
421 closed canopy forested ecosystems, such as the montane zone of Kosciuszko National Park,
422 make visual detection and certain forms of inter/intra specific cues difficult [8, 24].
423 Consequently, it was predicted that during summer greater olfactory cues owing to increased
424 temperatures, and increased invertebrate and microbial scavenging activity, would facilitate
425 shorter time to first detection of carcasses than during the other seasons [3, 5, 24, 82, 83].
426 However, there was no evidence to support this hypothesis.

427 Instead, it is possible that the same mechanism expected to facilitate shorter time to first
428 detection of carcasses, invertebrate and microbial scavenger activity, in fact, hindered detection
429 of carcasses by the vertebrate scavengers. During summer (warmer months) carcasses are

430 rapidly colonised (within minutes) *en masse* by invertebrates, and presumably microbes [83-
431 88]. The intense scavenging activity that follows accelerates carcass decay through the different
432 stages of decomposition, potentially at such a rate that vertebrate scavengers were given too
433 little time to detect carcasses i.e., the invertebrate scavengers outcompeted vertebrate
434 scavengers [83, 87, 88]. Indeed, previous observations of carcass persistence times in the study
435 site indicated that eastern grey kangaroo carcasses take at least twice as long to reach the dry
436 decay stage (only skin and bones remaining) in cool compared to warmer periods [46]. During
437 the winter monitoring period of this study, some carcasses did not reach the dry decay stage,
438 even after 60 days.

439 Many global studies have demonstrated that vertebrate scavenging rates are lowest during
440 summer when other food sources are more abundant and vice versa during winter [26-28]. In
441 our study, brushtail possums accounted for 81% of all recorded events during winter and they
442 scavenged three times more often during winter than during summer. Generally, the diet of
443 brushtail possums consists of leaves, flowers, fruit, (*Eucalyptus* and *Acacia*) and insects [89-
444 91], most of which are only seasonally available during warmer months. Therefore, this
445 marsupial may be exhibiting a dependence on carrion during winter that is similar to that of
446 other scavengers in northern hemisphere ecosystems [26-28]. The dependence of the brushtail
447 possum on carrion during winter likely influenced our analyses of the scavenger activity
448 response variable which recorded scavenging to be 2.789 times significantly more likely than
449 investigation of carcasses during winter when compared to summer. This may also explain why
450 carcasses took six times longer to be detected during summer (144 hours – 6 days) than during
451 winter (24 hours), contrary to our prediction that time to first arrival would be shortest in
452 summer. Collectively, during autumn, winter, and spring, 93% of the first detections of a
453 carcass were by either brushtail possums or raven spp. Conversely, during summer, only 57%
454 of the first detections of a carcass were by either brushtail possums or raven spp. Therefore, it

455 is possible that the brushtail possums and raven spp. may have a disproportionately greater
456 bearing on the time it takes the collective scavenger guild to first detect a carcass.

457 The species-specific breeding seasons of scavengers can also have profound impacts on their
458 respective scavenging behaviours [57, 64, 92, 93]. The Australian raven and little raven breed
459 from late winter into spring [94-96]. Initially, nest construction is prioritised in this early
460 breeding season, and the associated activities are characterised by frequent and short visitations
461 between the nest and sources of nesting material [97]. In our study, numerous remote camera
462 images captured during the spring monitoring period recorded raven spp. collecting hair and/or
463 fur from the carcasses, presumably for nest construction (Figure S1). Following nest
464 construction, chick rearing often requires breeding pairs to divide time between foraging,
465 feeding chicks, and being vigilant and protective of the nest [63]. These considerations often
466 mean that frequent but brief carcass visitations continue into the chick rearing season as the
467 breeding pairs frequently fly back and forth between the nest and food sources [63]. Inherently,
468 during this time both raven spp. incur greater energy costs associated with these activities, and
469 thus, must supplement their diets with protein rich sources and/or greater quantities of food,
470 such as carrion [57, 93]. Of all recorded raven spp. scavenging events during this study, 67%
471 were during spring, and this suggests that raven spp. may heavily rely on carrion to supplement
472 their diet, and that of their chicks, during the breeding season [57]. Further still, this flurry of
473 raven spp. scavenging during spring (a 170% increase on average annual raven spp. scavenging
474 events) accounted for 73% of all scavenging events for the collective scavenger guild.
475 Therefore, given their dominance during spring, it is likely that the frequent but brief
476 scavenging events, characteristic of the raven spp. breeding season, were also deterministic of
477 the significantly shorter scavenging event duration recorded for the collective scavenger guild
478 during spring.

479 The findings here regarding the raven spp. are also indirectly linked to the initial prediction for
480 longer scavenging event duration during winter – models dictated that scavenging event
481 duration during spring was 1.493 times significantly shorter than during winter. Whilst raven
482 spp. scavenging behaviours during their breeding seasons (spring) likely determined this result
483 and supported our prediction, it is juxtaposed to the initially used supporting evidence. That
484 being, many previous studies, mostly undertaken in the northern hemisphere, having
485 demonstrated scavenging rates to increase during winter in response to a lack of other available
486 food sources [26-28]. This raises several questions that need be addressed regarding the degree
487 to which the Australian alpine winter impacts food sources, species diets, and associated flow-
488 on effects to scavenging dynamics vs other alpine areas in the world, and indeed non-alpine
489 ecosystems that experience harsher winters.

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499 **Conclusion**

500 Scavenging dynamics in this study were highly seasonal, but dictated by the scavenging
501 activities and behaviours of mesoscavengers – the brushtail possum and raven spp.. The high
502 rate of scavenging by these species drove the seasonal trends in scavenging dynamics, but the
503 direction in which the seasonal effects impacted the scavenging response variables was not
504 always as predicted. This exemplifies the unexpected influence that seasons can have on
505 ecological processes linked to scavenging and highlights the need for seasonally replicated
506 experimental approaches in field-based scavenging research; the primary motivation for
507 undertaking this study.

508 The high rates of scavenging by the brushtail possum suggest that the species dietary status be
509 reconsidered, and with regards to the raven spp. a potential avenue for future work could be to
510 investigate the impact of carrion availability on breeding success. These findings, and indeed
511 such recommended future research (discussed in further detail in *Chapter 4 – Conclusion and*
512 *future directions*), have the potential to be of continental relevance given that the brushtail
513 possum is the most widespread Australian marsupial and that both raven spp. are also relatively
514 abundant across southern-east Australia [98].

515 In recent decades increased recognition of the ecosystem sustaining processes that scavengers
516 provide have advanced our understanding of scavenging dynamics, a previously misunderstood
517 and underappreciated area of ecology [7, 99]. In order to ensure that such scientific advances
518 are maintained, seasonal effects need to be accounted for in the field of scavenging ecology.
519 Not only are the findings of this study ecologically relevant to scavenging ecology within
520 Australia, but also serve as a more practical example of a robust and repeatable method for
521 monitoring and assessing scavenging dynamics surrounding carrion within any given
522 ecosystem.

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Chapter 2 – Supplementary Material



Figure 1. A raven spp. (*Corvus coronoides* or *Corvus mellori*) breeding pair collecting hair/fur from a kangaroo carcass during the spring monitoring period, presumably for nest construction.

Table 1. Moran's I measure of spatial independence for the various scavenging response variables.

Observed	Expected	SD	<i>p</i>
Species Richness Investigation Events			
0.006	-0.018	0.033	0.473
Species Richness Scavenging Events			
-0.026	-0.018	0.034	0.808
Time to First Detection			
-0.021	-0.018	0.032	0.918
Time to First Scavenging			
0.005	-0.018	0.033	0.490
Total Investigation Time			
0.009	-0.018	0.033	0.417
Average Investigation Time			
-0.012	-0.018	0.034	0.870
Total Scavenging Time			
-0.064	-0.018	0.032	0.150
Average Scavenging Time			
-0.040	-0.018	0.033	0.503

Table 2. The permutational multivariate analysis of variance (PERMANOVA) for investigation events species composition (top) and scavenging events species composition (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

	Df	Sums of sqs	Mean sqs	F Model	R²	<i>p</i>
Investigation Events						
Seasons	3	0.931	0.310	2.898	0.139	0.002
Residuals	54	5.779	0.107		0.861	
Total	57	6.710			1.000	
Scavenging Events						
Seasons	3	1.488	0.496	2.991	0.152	0.004
Residuals	50	8.289	0.166		0.848	
Total	53	9.776			1.000	

Table 3. Tukey’s honest significance tests between each of the seasons for the permutational multivariate analysis of variance (PERMANOVA) for investigation events species composition (top) and scavenging events species composition (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Contrast	F Model	R²	<i>p</i>	<i>p</i> adjusted
Investigation Events				
Autumn - Spring	4.554	0.140	0.005	0.030
Autumn - Summer	1.984	0.068	0.119	0.714
Autumn - Winter	1.658	0.058	0.169	1.000
Spring - Summer	1.147	0.041	0.358	1.000
Spring - Winter	4.995	0.156	0.003	0.018
Summer - Winter	3.436	0.117	0.015	0.090
Scavenging Events				
Autumn - Spring	3.252	0.115	0.009	0.054
Autumn - Summer	1.464	0.060	0.212	1.000
Autumn - Winter	3.857	0.138	0.006	0.036
Spring - Summer	2.974	0.103	0.030	0.180
Spring - Winter	5.500	0.169	0.002	0.012
Summer - Winter	1.797	0.067	0.152	0.912

Table 4. The analysis of similarities (ANOSIM) for investigation events species composition (top) and scavenging events species composition (bottom). Significant p-values ($p < 0.05$) are denoted in bold.

ANOSIM statistic R	Significance
Investigation Events	
0.109	0.001
Scavenging Events	
0.109	0.002

Table 5. The similarity percentages (SIMPER) analysis demonstrating the contribution of each species to explaining differences between the seasons in investigation event species composition.

	Average	SD	Ratio	AVA	AVB	CUMSUM
Summer vs Autumn						
Brushtail possum	0.433	0.244	1.772	47.214	27.067	0.705
Raven spp.	0.039	0.046	0.849	1.500	2.333	0.769
Feral cat	0.031	0.074	0.420	0.357	1.933	0.819
Dingo	0.030	0.046	0.644	1.571	0.667	0.868
Feral pig	0.022	0.043	0.511	0.214	1.200	0.904
Pied currawong	0.022	0.035	0.606	0.786	1.000	0.939
Wedge-tailed eagle	0.018	0.034	0.538	0.429	0.800	0.969
Red fox	0.017	0.033	0.504	0.000	1.400	0.996
Spotted-tail quoll	0.002	0.007	0.336	0.000	0.133	1.000
Summer vs Winter						
Brushtail possum	0.436	0.247	1.765	47.214	44.929	0.781
Raven spp.	0.039	0.054	0.721	1.500	2.857	0.850
Pied currawong	0.026	0.042	0.626	0.786	1.786	0.897
Dingo	0.021	0.033	0.651	1.571	0.071	0.936
Red fox	0.012	0.020	0.593	0.000	0.857	0.957
Wedge-tailed eagle	0.008	0.023	0.338	0.429	0.214	0.971
Feral cat	0.007	0.011	0.651	0.357	0.429	0.984
Feral pig	0.006	0.011	0.541	0.214	0.357	0.995
Spotted-tail quoll	0.003	0.010	0.305	0.000	0.286	1.000

Summer vs Spring						
Brushtail possum	0.375	0.223	1.683	47.214	23.600	0.586
Raven spp.	0.126	0.169	0.743	1.500	13.200	0.783
Pied currawong	0.089	0.146	0.607	0.786	6.200	0.921
Dingo	0.034	0.045	0.764	1.571	1.800	0.974
Wedge-tailed eagle	0.006	0.022	0.269	0.429	0.000	0.984
Feral cat	0.005	0.008	0.685	0.357	0.267	0.992
Feral pig	0.005	0.011	0.417	0.214	0.067	0.999
Red fox	0.001	0.003	0.248	0.000	0.067	1.000
Spotted-tail quoll	0.000	0.000	0.000	0.000	0.000	1.000
Autumn vs Winter						
Brushtail possum	0.347	0.223	1.559	27.067	44.929	0.682
Raven spp.	0.047	0.055	0.865	2.333	2.857	0.775
Pied currawong	0.027	0.041	0.646	1.000	1.786	0.827
Feral cat	0.026	0.058	0.454	1.933	0.429	0.879
Red fox	0.021	0.027	0.794	1.400	0.857	0.921
Feral pig	0.017	0.031	0.549	1.200	0.357	0.954
Wedge-tailed eagle	0.011	0.019	0.607	0.800	0.214	0.976
Dingo	0.007	0.014	0.528	0.667	0.071	0.991
Spotted-tail quoll	0.005	0.010	0.457	0.133	0.286	1.000
Autumn vs Spring						
Brushtail possum	0.250	0.174	1.442	27.067	23.600	0.430
Raven spp.	0.138	0.167	0.826	2.333	13.200	0.667
Pied currawong	0.094	0.146	0.641	1.000	6.200	0.827
Dingo	0.030	0.047	0.647	0.667	1.800	0.879

Feral cat	0.026	0.060	0.436	1.933	0.267	0.924
Feral pig	0.017	0.031	0.548	1.200	0.067	0.953
Red fox	0.015	0.028	0.552	1.400	0.067	0.979
Wedge-tailed eagle	0.010	0.018	0.568	0.800	0.000	0.997
Spotted-tail quoll	0.002	0.005	0.377	0.133	0.000	1.000

Winter vs Spring

Brushtail possum	0.272	0.192	1.419	44.929	23.600	0.524
Raven spp.	0.119	0.142	0.840	2.857	13.200	0.754
Pied currawong	0.080	0.115	0.689	1.786	6.200	0.908
Dingo	0.023	0.039	0.583	0.071	1.800	0.952
Red fox	0.011	0.016	0.668	0.857	0.067	0.973
Feral cat	0.006	0.009	0.689	0.429	0.267	0.984
Feral pig	0.004	0.005	0.659	0.357	0.067	0.991
Spotted-tail quoll	0.003	0.008	0.326	0.286	0.000	0.996
Wedge-tailed eagle	0.002	0.008	0.273	0.214	0.000	1.000

Table 6. The similarity percentages (SIMPER) analysis demonstrating the contribution of each species to explaining differences between the seasons in scavenging event species composition.

	Average	SD	Ratio	AVA	AVB	CUMSUM
Summer vs Autumn						
Brushtail possum	0.355	0.293	1.214	30.615	25.500	0.474
Raven spp.	0.238	0.211	1.129	5.154	20.083	0.793
Feral pig	0.084	0.137	0.610	1.846	4.000	0.904
Wedge-tailed eagle	0.039	0.086	0.454	0.769	1.750	0.956
Red fox	0.017	0.025	0.702	0.000	1.167	0.980
Dingo	0.010	0.025	0.388	0.385	0.333	0.992
Pied currawong	0.006	0.022	0.261	0.231	0.000	1.000
Feral cat	0.000	0.000	0.000	0.000	0.000	1.000
Spotted-tail quoll	0.000	0.000	0.000	0.000	0.000	1.000
Summer vs Winter						
Brushtail possum	0.496	0.290	1.711	30.615	86.357	0.694
Raven spp.	0.135	0.154	0.874	5.154	14.214	0.883
Feral pig	0.030	0.072	0.408	1.846	0.714	0.924
Wedge-tailed eagle	0.026	0.071	0.362	0.769	4.429	0.960
Pied currawong	0.017	0.050	0.345	0.231	4.000	0.985
Dingo	0.004	0.014	0.311	0.385	0.000	0.991
Red fox	0.003	0.009	0.293	0.000	0.214	0.995
Spotted-tail quoll	0.002	0.009	0.292	0.000	0.286	0.998
Feral cat	0.001	0.006	0.253	0.000	0.071	1.000
Summer vs Spring						

Raven spp.	0.410	0.310	1.322	5.154	71.000	0.500
Brushtail possum	0.252	0.254	0.993	30.615	14.600	0.807
Pied currawong	0.091	0.162	0.564	0.231	8.133	0.919
Feral pig	0.026	0.047	0.541	1.846	0.733	0.950
Dingo	0.025	0.041	0.608	0.385	2.267	0.980
Wedge-tailed eagle	0.010	0.037	0.283	0.769	0.133	0.993
Spotted-tail quoll	0.005	0.011	0.394	0.000	0.667	0.998
Red fox	0.002	0.006	0.254	0.000	0.133	1.000
Feral cat	0.000	0.000	0.000	0.000	0.000	1.000

Autumn vs Winter

Brushtail possum	0.440	0.281	1.564	25.500	86.357	0.628
Raven spp.	0.164	0.158	1.041	20.083	14.214	0.862
Feral pig	0.038	0.071	0.530	4.000	0.714	0.916
Wedge-tailed eagle	0.029	0.060	0.483	1.750	4.429	0.957
Pied currawong	0.014	0.048	0.291	0.000	4.000	0.977
Red fox	0.011	0.017	0.652	1.167	0.214	0.992
Spotted-tail quoll	0.002	0.008	0.289	0.000	0.286	0.996
Dingo	0.002	0.005	0.367	0.333	0.000	0.998
Feral cat	0.001	0.005	0.240	0.000	0.071	1.000

Autumn vs Spring

Raven sp.	0.365	0.276	1.325	20.083	71.000	0.497
Brushtail possum	0.205	0.219	0.939	25.500	14.600	0.776
Pied currawong	0.080	0.149	0.537	0.000	8.133	0.885
Feral pig	0.036	0.055	0.649	4.000	0.733	0.933
Dingo	0.021	0.037	0.565	0.333	2.267	0.962

Wedge-tailed eagle	0.014	0.023	0.598	1.750	0.133	0.980
Red Fox	0.010	0.015	0.691	1.167	0.133	0.994
Spotted-tail quoll	0.004	0.011	0.392	0.000	0.667	1.000
Feral cat	0.000	0.000	0.000	0.000	0.000	1.000
----- Winter vs Spring -----						
Brushtail possum	0.353	0.245	1.442	86.357	14.600	0.469
Raven spp.	0.290	0.248	1.166	14.214	71.000	0.853
Pied currawong	0.068	0.120	0.567	4.000	8.133	0.943
Dingo	0.015	0.030	0.510	0.000	2.267	0.963
Wedge-tailed eagle	0.014	0.049	0.290	4.429	0.133	0.982
Feral pig	0.006	0.009	0.614	0.714	0.733	0.990
Spotted-tail quoll	0.005	0.010	0.474	0.286	0.667	0.996
Red Fox	0.002	0.007	0.373	0.214	0.133	0.999
Feral cat	0.001	0.003	0.247	0.071	0.000	1.000

Table 7. The Akaike information criterion (AIC) ranking of each of the models for investigation events species richness (top) and scavenging events species richness (bottom). Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
Investigation Events Species Richness						
<i>1</i>	<i>Seasons</i>	<i>4</i>		<i>192.082</i>	<i>0.00</i>	<i>0.292</i>
<i>2</i>	<i>Null (intercept only)</i>	<i>1</i>		<i>192.131</i>	<i>0.05</i>	<i>0.285</i>
<i>3</i>	<i>Seasons + Altitude</i>	<i>5</i>		<i>193.360</i>	<i>1.28</i>	<i>0.154</i>
<i>4</i>	<i>Altitude</i>	<i>2</i>		<i>193.537</i>	<i>1.46</i>	<i>0.141</i>
Scavenging Events Species Richness						
<i>1</i>	<i>Null (intercept only)</i>	<i>1</i>		<i>194.288</i>	<i>0</i>	<i>0.400</i>
<i>2</i>	<i>Seasons</i>	<i>4</i>		<i>195.294</i>	<i>1.01</i>	<i>0.241</i>
<i>3</i>	<i>Altitude</i>	<i>2</i>		<i>195.546</i>	<i>1.26</i>	<i>0.213</i>
<i>4</i>	<i>Seasons + Altitude</i>	<i>5</i>		<i>196.441</i>	<i>2.15</i>	<i>0.136</i>

Table 8. The base generalised linear model (GLM) for investigation event species richness (top) and scavenging event species richness (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	p
Investigation Events				
Intercept	1.852	0.531	3.487	<0.001
Spring	-0.236	0.122	-1.933	0.053
Summer	-0.258	0.125	-2.055	0.040
Winter	-0.264	0.125	-2.104	0.035
Altitude	<0.001	<0.001	-0.853	0.394
Scavenging Events				
Intercept	1.498	0.779	1.922	0.055
Spring	0.293	0.179	1.640	0.101
Summer	-0.095	0.202	-0.470	0.638
Winter	0.174	0.186	0.934	0.350
Altitude	< - 0.001	<0.001	-0.920	0.358

Table 9. The Akaike information criterion (AIC) ranking of each of the models for time to first detection (top) and time to first scavenging (bottom). Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
Time to First Detection						
<i>1</i>	<i>Seasons</i>	<i>5</i>	<i>-282.223</i>	<i>575.6</i>	<i>0.00</i>	<i>0.730</i>
<i>2</i>	<i>Seasons + Altitude</i>	<i>7</i>	<i>-279.511</i>	<i>577.6</i>	<i>1.99</i>	<i>0.269</i>
3	Null (intercept only)	1	-292.747	589.7	14.11	0.001
4	Altitude	4	-291.294	592.5	16.92	0.000
Time to First Scavenging						
<i>1</i>	<i>Seasons</i>	<i>5</i>	<i>-330.042</i>	<i>671.3</i>	<i>0.00</i>	<i>0.438</i>
<i>2</i>	<i>Null (intercept only)</i>	<i>2</i>	<i>-333.864</i>	<i>672.0</i>	<i>0.65</i>	<i>0.317</i>
<i>3</i>	<i>Seasons + Altitude</i>	<i>9</i>	<i>-325.495</i>	<i>673.2</i>	<i>1.93</i>	<i>0.167</i>
4	Altitude	5	-331.655	674.8	3.46	0.078

Table 10. The base generalised linear model (GLM) for time to first detection (top) and time to first scavenging (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	<i>p</i>
Time to First Detection				
Intercept	3.796	0.260	14.621	< 0.001
Spring	-0.308	0.369	-0.835	0.404
Summer	1.148	0.373	3.077	0.002
Winter	-0.729	0.374	-1.948	0.051
Variable	edf	Ref.df	Chi.sq	<i>p</i>
Altitude	2.265	2.867	5.16	0.13
Time to First Scavenging				
Intercept	5.255	0.281	18.674	< 0.001
Spring	-0.650	0.387	-1.681	0.092
Summer	0.354	0.397	0.893	0.372
Winter	-0.808	0.392	-2.064	0.039
Variable	edf	Ref.df	Chi.sq	<i>p</i>
Altitude	3.294	4.115	9.227	0.048

Table 11. The Tukey’s honest significance test of the base generalised linear model (GLM) for time to first detection (top) and time to first scavenging (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Contrast	Estimate	SE	df	t ratio	p
Time to First Detection					
Autumn - Spring	0.308	0.448	51.7	0.688	0.901
Autumn - Summer	-1.148	0.453	51.7	-2.535	0.066
Autumn - Winter	0.729	0.454	51.7	1.605	0.385
Spring - Summer	-1.456	0.453	51.7	-3.216	0.012
Spring - Winter	0.421	0.456	51.7	0.922	0.793
Summer - Winter	1.876	0.461	51.7	4.071	< 0.001
Time to First Scavenging					
Autumn - Spring	0.650	0.506	47.7	1.286	0.576
Autumn - Summer	-0.354	0.519	47.7	-0.683	0.903
Autumn - Winter	0.808	0.512	47.7	1.578	0.400
Spring - Summer	-1.004	0.499	47.7	-2.013	0.198
Spring - Winter	0.158	0.494	47.7	0.320	0.989
Summer - Winter	1.163	0.507	47.7	2.291	0.114

Table 12. The Akaike information criterion (AIC) ranking of each of the models for scavenging activity. Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
<i>1</i>	<i>Seasons</i>	<i>5</i>	<i>-4279.252</i>	<i>8568.5</i>	<i>0.00</i>	<i>0.690</i>
<i>2</i>	<i>Seasons + Altitude</i>	<i>6</i>	<i>-4279.054</i>	<i>8570.1</i>	<i>1.61</i>	<i>0.309</i>
3	Null (intercept only)	2	-4289.519	8583.0	14.53	0.000
4	Altitude	3	-4289.413	8584.8	16.32	0.000

Table 13. The base generalised linear model (GLM) for scavenging activity. Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	z value	p
Spring	0.776	0.260	2.981	0.003
Summer	-0.358	0.274	-1.307	0.191
Winter	0.667	0.264	2.529	0.011
Altitude	-0.059	0.093	-0.628	0.530

Table 14. The Tukey's honest significance tests of the base generalised linear model (GLM) for scavenging activity. Significant p-values ($p < 0.05$) are denoted in **bold**.

Contrast	Estimate	SE	df	z ratio	p
Autumn - Spring	-0.776	0.260	Inf	-2.981	0.015
Autumn - Summer	0.358	0.274	Inf	1.307	0.558
Autumn - Winter	-0.667	0.264	Inf	-2.529	0.056
Spring - Summer	1.134	0.265	Inf	4.278	< 0.001
Spring - Winter	0.109	0.255	Inf	0.429	0.974
Summer - Winter	-1.025	0.269	Inf	-3.813	< 0.001

Table 15. The Akaike information criterion (AIC) ranking of each of the models for investigation event duration (top) and scavenging event duration (bottom). Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
Investigation Event Duration						
<i>1</i>	<i>Null (intercept only)</i>	<i>3</i>	<i>-2567.223</i>	<i>5140.5</i>	<i>0.00</i>	<i>0.526</i>
<i>2</i>	<i>Seasons</i>	<i>6</i>	<i>-2565.157</i>	<i>5142.3</i>	<i>1.89</i>	<i>0.204</i>
<i>3</i>	<i>Altitude</i>	<i>4</i>	<i>-2567.223</i>	<i>5142.5</i>	<i>2.01</i>	<i>0.193</i>
<i>4</i>	<i>Seasons + Altitude</i>	<i>7</i>	<i>-2565.140</i>	<i>5144.3</i>	<i>3.87</i>	<i>0.076</i>
Scavenging Event Duration						
<i>1</i>	<i>Seasons</i>	<i>6</i>	<i>-13666.05</i>	<i>27344.1</i>	<i>0.00</i>	<i>0.592</i>
<i>2</i>	<i>Seasons + Altitude</i>	<i>7</i>	<i>-13665.47</i>	<i>27345.0</i>	<i>0.86</i>	<i>0.385</i>
<i>3</i>	<i>Null (intercept only)</i>	<i>3</i>	<i>-13672.62</i>	<i>27351.3</i>	<i>7.14</i>	<i>0.017</i>
<i>4</i>	<i>Altitude</i>	<i>4</i>	<i>-13672.57</i>	<i>27353.2</i>	<i>9.04</i>	<i>0.006</i>

Table 16. The base generalised linear model (GLM) for investigation event duration (top) and scavenging event duration (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	p
Investigation Event Duration				
Intercept	-0.045	0.151	-0.298	0.765
Spring	-0.027	0.207	-0.130	0.897
Summer	0.357	0.216	1.657	0.098
Winter	0.196	0.209	0.937	0.349
Altitude	0.014	0.075	0.182	0.856
Scavenging Event Duration				
Intercept	2.683	0.124	21.678	< 0.001
Spring	-0.639	0.163	-3.925	< 0.001
Summer	-0.257	0.179	-1.439	0.150
Winter	-0.238	0.165	-1.445	0.149
Altitude	0.063	0.059	1.077	0.282

Table 17. The Tukey’s honest significance test of the base generalised linear model (GLM) for investigation event duration (top) and scavenging event duration (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Contrast	Estimate	SE	df	z ratio	<i>p</i>
Investigation Event Duration					
Autumn - Spring	0.027	0.207	Inf	0.130	1.000
Autumn - Summer	-0.357	0.216	Inf	-1.657	0.347
Autumn - Winter	-0.196	0.209	Inf	-0.937	0.785
Spring - Summer	-0.384	0.210	Inf	-1.825	0.261
Spring - Winter	-0.223	0.204	Inf	-1.093	0.694
Summer - Winter	0.161	0.213	Inf	0.755	0.874
Scavenging Event Duration					
Autumn - Spring	0.639	0.163	Inf	3.925	< 0.001
Autumn - Summer	0.257	0.179	Inf	1.439	0.475
Autumn - Winter	0.238	0.165	Inf	1.445	0.472
Spring - Summer	-0.382	0.167	Inf	-2.183	0.102
Spring - Winter	-0.401	0.152	Inf	-2.643	0.041
Summer - Winter	-0.019	0.169	Inf	-0.112	1.000

1 Chapter 3 – Absence of apex predator top-down

2 effects on mesoscavengers

3 Abstract

4 Apex predators are renowned for their abilities to structure ecological communities and
5 regulate ecosystems. Conventional approaches to ecological theories regarding apex predators
6 often only consider their predatory capabilities. However, despite their name, most apex
7 predators scavenge and whilst largely opportunistic, scavenging rates can be substantial in
8 response to low prey availability. Consequently, within a scavenging context, predators can
9 influence the scavenging behaviours of other species as well as scavenging dynamics more
10 broadly. Further still, scavenger guilds are composed of a unique and diverse group of taxa,
11 and thus, the top-down scavenging effects of apex predators have the potential to result in novel
12 inter-specific interactions surrounding carrion. Despite the presence of and scavenging by an
13 apex predator, the dingo (*Canis dingo*), scavenging dynamics within Australian Alps have been
14 demonstrated to be dominated by three low ranking mesoscavengers, the brushtail possum
15 (*Trichosurus vulpecula*) and two raven species (*Corvus coronoides* and *Corvus mellori*) –
16 Chapter 2. Therefore, here, top-down scavenging effects exerted by the dingo were assessed
17 regarding potential impacts on the scavenging activities of these mesoscavengers. Overall,
18 dingoes were found to scavenge at low rates and no evidence was found to suggest dingoes
19 influenced the scavenging behaviours of brushtail possums or raven spp. This raised several
20 questions regarding the extent to which dingoes scavenge when anthropogenically suppressed,
21 and in response to an overabundance of carrion. Each of these concepts are discussed herein,
22 with a particular focus on density dependent predator theories, and how each scenario may
23 determine the extent to which dingoes can exert top-down scavenging effects.

24 **Introduction**

25 Globally, apex predators are one of the most recognised and studied groups of taxa due to their
26 status as keystone species i.e., species that drive community dynamics and structure food webs
27 within ecosystems [1]. The predatory activities of such species can directly regulate prey
28 populations and their recognition as potentially dangerous predators can indirectly influence
29 movement and habitat use by other species [2]. Both of these mechanisms facilitate trophic
30 cascades which can maintain biodiversity in both the animal and plant kingdoms [3, 4]. These
31 types of cascades are present in terrestrial, aquatic, and marine ecosystems. Some well cited
32 examples include the suppression of elk (*Cervus canadensis*) populations following the
33 reintroduction of grey wolves (*Canis lupus*) to Yellowstone National Park [5, 6], and the
34 maintenance of near shore community biodiversity in the Aleutian Islands (Alaska) by sea
35 otters (*Enhydra lutris*) [7].

36 Predators are typically classified as either an apex predator or mesopredator. The former are
37 species characterised by their position at the top of the food chain and lack of natural predators
38 [8], whilst the later are “any midranking predator in a food web, regardless of its size or
39 taxonomy” [9]. The presence or absence of either species group within ecosystems can cause
40 trophic cascades that can result from their interspecific interactions or lack thereof [8]. The
41 mesopredator release hypothesis explains how constraints to the population growth of
42 mesopredators, which are controlled by competitively dominant apex predators, are released
43 following a decline in apex predator population levels [10-12]. Theories such as this, and
44 indeed others, demonstrate how the preservation of apex predators alone, because of their high
45 trophic level, can have flow-on effects on ecosystem health, biodiversity, and the conservation
46 of other subordinate species [13].

47 Most conventional approaches to ecological theories regarding apex predators, however, often
48 only consider their role as predatory species [2]. Whilst this predatory role is undoubtedly
49 crucial to a plethora of essential ecosystem services globally, the other roles that apex predators
50 play are often overlooked [14, 15]. Despite their name, apex predators are not purely predatory,
51 most, if not all, scavenge on dead animal remains (carrion) at some point in their lifetime. This
52 phenomenon is mostly opportunistic, however, scavenging rates can be substantial in response
53 to low prey availability [16]. For example, in Scandinavia, wolverine (*Gulo gulo*) feeding
54 strategies (i.e., predation vs scavenging) were demonstrated to be seasonally dependent on prey
55 body condition and carrion supply which in combination promoted predation during summer
56 and scavenging during winter [17]. This flexibility in diet is important to distinguish, especially
57 so regarding predators, because there is a clear differentiation in the fitness outcomes of
58 scavenging which results in only a fraction of the energy costs associated with predation [18,
59 19]. Although only recently considered, it is, therefore, evident that scavenging may provide
60 an important energy pathway for predators. Not only does this have cause to force the
61 reevaluation of the diets and energy budgets of predators but also expands the plethora of
62 ecosystem processes dictated by predators to include scavenging.

63 Apex predators have most often been credited with dictating scavenging dynamics within
64 ecosystems via the provision of carrion from their partially consumed prey [18]. Whilst this is
65 an important consideration, it overshadows the fact that predators can also dictate scavenging
66 dynamics via their own scavenging activities surrounding non-prey killed carrion [18]. Apex
67 predators, similarly to vultures (obligate scavengers), can rapidly consume carrion biomass,
68 including bones, and their presence at carrion can, therefore, accelerate carcass decomposition
69 [20]. Scavenging on non-prey killed carrion by dominant scavenging predators can also directly
70 force other scavengers from carrion (competition) and/or influence how they access carrion
71 (landscape of fear) [19, 21-26]. For instance, the black bear (*Ursus americanus*), was

72 demonstrated to influence the structure and composition of its scavenger guild in northern
73 California [25]. Specifically, the presence of black bears at carcasses limited scavenger species
74 richness, significantly reduced sum scavenging time by other scavengers, and increased the
75 nestedness of the scavenger guild [25]. These types of top-down effects are often extreme
76 enough that many mesoscavengers utilise spatial and temporal measures, as well as flexible
77 behavioural strategies, to avoid confrontations with competitively dominant apex scavengers
78 and exploit carrion [25, 27].

79 Within a scavenger guild a unique selection of species are present [28]. There is potential for
80 dominant scavenging predators to influence the scavenging behaviours and activities of any of
81 these species, be it directly or indirectly. Consequently, a suite of novel interspecific
82 interactions can occur that are not purely considered competitive interactions strictly between
83 predators, as discussed above and more broadly in the literature. The ‘mesoscavenger release
84 hypothesis’ was recently coined and is essentially a reframing of the ‘mesopredator release
85 hypothesis’ within a scavenging context [21]. Here, the importance of recognising predators
86 also as scavengers is emphasised because the dynamics of such a ‘mesoscavenger release’ are
87 equally as complex as the conventional ‘mesopredator release’ and have the potential to impact
88 a much wider breadth of taxa [22, 25, 26, 28, 29]. For instance, in Tasmania (an island state of
89 Australia), recent disease related declines in the population of Tasmanian devils (*Sarcophilus*
90 *harrisii*), Tasmania’s native marsupial apex predator and scavenger, has had cascading effects
91 on other scavengers and scavenging dynamics [26, 29]. As a result, in the areas where
92 Tasmanian devil density is low, a native avian scavenger, the forest raven (*Corvus tasmanicus*),
93 scavenged longer on carcasses, whilst an invasive mesopredator, the feral cat (*Felis catus*),
94 scavenged at more carcasses. This mesoscavenger release highlighted how apex predators
95 within a scavenging context can exert top-down effects on a novel selection of species only
96 present within scavenger guilds.

97 The Tasmanian devil is, however, absent from mainland Australia. The dingo (*Canis dingo*),
98 despite having a brief evolutionary and ecological history in Australia (~ 4000 years) [30], has
99 established itself as mainland Australia's dominant apex predator. Many government agencies
100 and agricultural communities considering the dingo a pest species. However, the dingo can
101 have a strong regulatory role within many Australian ecosystems that spans multiple trophic
102 levels and includes the beneficial control of overabundant prey (i.e., kangaroo –
103 *Macropodidae*) [31, 32] and the maintenance of biodiversity via suppression of invasive
104 mesopredators (i.e., red fox – *Vulpes vulpes* and feral cat – *Felius catus*) [30, 33-36]. The
105 regulatory role of the dingo is not only localised to specific ecosystems either, but is also
106 exerted at a continental scale [37]. Whilst there has been a strong focus on the predation effects
107 of dingoes, a common trend in the global predator-based literature, the dingo is also an
108 extremely effective scavenger [38-41]. Dingoes, like most predators, are facultative scavengers
109 and will opportunistically scavenge on carrion when available [38], however, carrion can
110 become a crucial part of their diet during food shortages [41].

111 Numerous studies have documented the scavenging activities of dingoes [38, 41-43], the most
112 comprehensive of which experimentally monitored dingo scavenging across three Australian
113 ecosystems (alpine, forest, and desert) [38]. This study by Spencer and Newsome (2021) found
114 that dingo scavenging dynamics were highly variable and complex, but intrinsically linked to
115 the seasons, and that in certain contexts dingoes could substantially contribute to carrion
116 biomass loss. However, Spencer and Newsome (2021) did note a limitation of their study being
117 that the top-down scavenging effects of dingoes were not assessed, and thus, emphasised that
118 additional seasonal replications may yield more detailed information on the role of dingoes as
119 dominant scavenging predators. Of relevance here is Forsyth et al.'s 2014 study that found
120 dingoes to limit the spatial and temporal availability of carcasses to an invasive mesoscavenger,
121 the red fox [40]. Also of relevance, but to a lesser degree, are the findings of Schlacher et al.

122 (2014) which demonstrated domestic dogs (*Canis lupus familiaris*), close relatives of the dingo,
123 to outcompete native scavengers for carrion on urban beaches [44]. However, these are the only
124 two such studies describing the top-down scavenging effects of dingoes, and indeed, a recent
125 review of canid scavenging effects by Wirsing and Newsome (2020) found that the scavenging
126 effects of most canid species remain little understood [39]. Furthermore, the top-down effects
127 of dingoes in general (i.e., those not limited to scavenging) are often only considered regarding
128 larger mesopredators and/or competitors [30, 33-37, 45]. This makes it difficult to transfer such
129 findings to a scavenging context given the diverse selection of species present within scavenger
130 guilds. Therefore, the fear effects that dingoes exert surrounding carrion and its impacts on
131 more novel scavenger species (i.e., potential prey) is an area that warrants further attention.

132 The brushtail possum (*Trichosurus vulpecula*) and raven spp. (*Corvus coronoides* and *Corvus*
133 *mellori*) are such facultative scavengers that could potentially have novel interspecific
134 interactions with dingoes surrounding carrion. As outlined in *Chapter 2*, these species were the
135 most abundant scavengers observed in the Australian Alps and dominated the scavenging
136 dynamics of the guild. When scavenging, these species must consider the threats posed to them
137 by dominant scavenging predators such as the dingo. A key question is, therefore, whether the
138 high rates of scavenging by brushtail possums and raven spp. are regulated by the top-down
139 scavenging effects of dingoes. To understand if such a scavenging dynamic was important
140 within the Australian Alps, firstly, the baseline scavenging rates of the dingo at two temporal
141 scales (daily and seasonal) was determined. By understanding the ways in which an apex
142 predator utilises an opportunistic resource, the extent to which dingo scavenging effects impact
143 the scavenging of other species can then be established. Specifically, the presence of dingoes
144 at carcasses, whether it be to investigate or scavenge, was analysed to determine how it
145 influences the probability and duration of scavenging by brushtail possums and raven spp..
146 Broadly, it was predicted that predator avoidance (i.e., the landscape of fear [46]) behaviours

147 would determine the scavenging dynamics of brushtail possums and raven spp. surrounding
148 carrion [24]. Specifically, it was expected that at carcasses visited by dingoes, brushtail
149 possums would investigate carcasses more often than scavenging on the carcass itself. In
150 addition, it was also predicted that at carcasses visited by dingoes, brushtail possum scavenging
151 bouts would be of shorter duration. Similar results were expected for raven spp., however, to a
152 lesser degree given their more effective predator avoidance abilities (flight), and thus, lower
153 perceived risk to predation. This study provides insights into the role an apex predator has
154 within a scavenger guild and uncovers the strength of intraspecific interactions between
155 predator and novel prey occurring around carrion.

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166 **Methods**

167 Due to both *Chapters 2* and *3* being formatted for submission to relevant scientific journals,
168 there is unavoidable repetition of methods between the two chapters.

169 **Study site**

170 This study was conducted in Kosciuszko National Park, located in southern New South Wales,
171 Australia. This region includes many of the highest peaks and ranges within Australia's Great
172 Dividing Range, including Australia's highest peak – Mount Kosciuszko. The landscape is
173 delineated into three altitudinally distinct ecological communities – montane (500m-1500m)
174 subalpine (1500m-1850m), and alpine (1850m-2228m) [47]. This work was undertaken within
175 the montane zone (between approximately 1000m – 1500m) which is characterised by forest
176 stands dominated by snow gum (*Eucalyptus pauciflora*) in association with various other
177 *Eucalyptus* species.

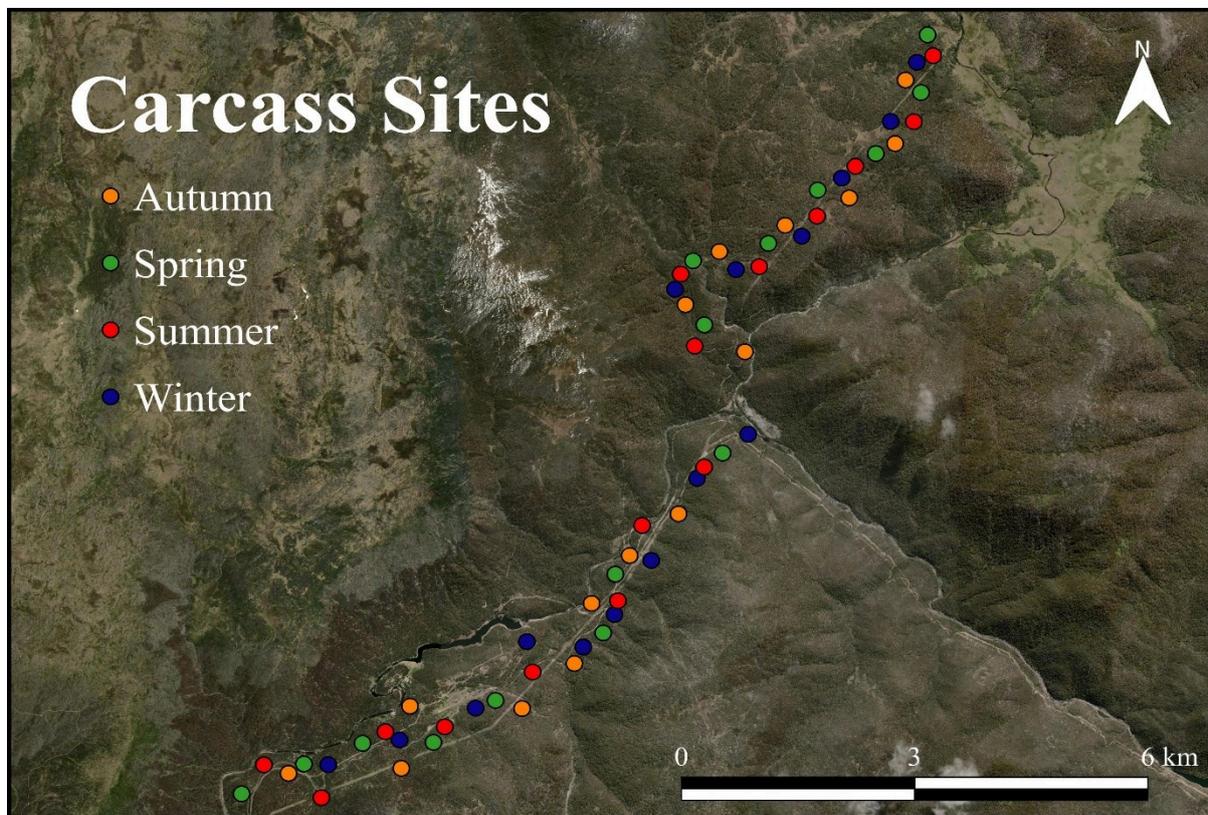
178 **Ethics, licenses, and permits**

179 The following described work received all required ethics, licenses, and permits approved by
180 the relevant authorities (i.e., The University of Sydney; New South Wales Office of
181 Environment and Heritage; and New South Wales National Parks and Wildlife Services). All
182 kangaroo carcasses used for the purposes of this research were sourced fresh and locally from
183 existing authorised and legally approved management culls that are conducted to control
184 overabundant kangaroo populations.

185 **Fieldwork**

186 A 15 km transect was established through Kosciuszko National Park along which all carcass
187 monitoring took place (Figure 1). This transect ran northeast – southwest from a border region
188 of the national park inwards towards its interior and was selected due to its accessibility (i.e.,

189 road access) and because it is a relatively undisturbed area with little human activity.
190 Monitoring periods were established to coincide with the four seasons: autumn – March 2020;
191 winter – July 2020; spring – October 2020; summer – January 2021. During each season, 15
192 sites were established along the transect, separated by approximately 1 km from the nearest
193 sites monitored within the same season and approximately 250m from the nearest sites
194 monitored during other seasons (60 different sites in total; Figure 1). The separation of the sites
195 as such ensured a level of spatial independence and it prevented habituation of scavengers to a
196 carrion source location [38, 48-50].



197 **Figure 1.** Location of the monitoring transect along which all 60 carcass sites were established
198 within Kosciuszko National Park (red = summer, orange = autumn, blue = winter, green =
199 spring).
200

201 Within each seasonal monitoring period, one fresh eastern grey kangaroo carcass was placed
202 at each site (15 total). The carcasses ranged in weight from 10kg-70kg and on average were
203 28.3kg (± 1.498 – standard deviation). Each carcass was secured, using wire ties, to star pickets
204 driven into the ground to ensure they remained in situ to be monitored for 60 days [38, 48, 49].
205 Vertebrate scavenger activity was monitored at each site using a Reconyx PC800 Hyperfire™
206 remote camera. Each camera was placed on a free-standing star picket three meters north of
207 the carcass – the southern aspect of the cameras decreased exposure to direct sunlight which
208 would otherwise reduce image quality. The cameras were calibrated to take photographs
209 continuously (approximately one image per second) when triggered by thermal movement (i.e.,
210 rapid-fire, no wait period). These approaches and methods follow those previously used in
211 field-based scavenging research [38, 48-54].

212 **Analyses**

213 Remote camera images were analysed for species presence and the number of individuals of a
214 species present. In order to determine distinct visitations of species, an ‘event’ was
215 characterised as a visitation by a species that occurred more than 10 minutes after the last
216 visitation by that same species. Only species-specific events could be characterised because
217 identification of individuals for most species was not possible. An event was characterised as
218 a ‘scavenging event’ if the species present scavenged on the monitored carcass in at least one
219 of the remote camera images consisting of that event, otherwise the event was characterised as
220 an ‘investigation event’. Data from species recorded to have scavenged at least once were
221 included in the statistical analyses.

222 The R software environment (version 1.4.1717) was used for all statistical analyses. Similarly
223 to *Chapter 2*, scavenger activity (i.e., probability of a scavenger investigating vs scavenging a
224 carcass; binomial distribution), and investigation and scavenging event duration (in minutes;

225 Gamma distribution) were again used as response variable. However, these variables were only
226 modelled species specifically for each of the scavengers in question – the dingo, brushtail
227 possum, and raven spp.. For the dingo, each of the response variables were modelled only
228 against the explanatory variable of season. For the brushtail possum and raven spp., each of the
229 response variables were modelled against the explanatory variables of season and dingo
230 presence. Dingo presence was a binary explanatory variable that simply recorded whether any
231 given monitored carcass was visited by a dingo (1) or not visited by a dingo (0). Altitude was
232 excluded from these models given that it was demonstrated to be insignificant in Chapter 2.
233 These models were constructed twice, once using only investigation events and once using
234 only scavenging events. Only one model was constructed for scavenger activity as it is a
235 binomial response variable (either investigation; 0, or scavenging; 1) designed specifically to
236 determine the probability of either an investigation event or scavenging event.

237 The relationships of each of these scavenging response variables with the explanatory variables
238 of season and dingo presence were modelled using either generalised linear models (GLM; R
239 Package ‘lme4’ [55]) or generalised linear mixed models (GLMM; R Package ‘lme4’ [55]),
240 and in the case of poorly fit models generalised additive models were utilised (GAM; R
241 Package ‘mgcv’ [56]). To determine the most parsimonious model(s), Akaike information
242 criterion (AIC) [57] was used ($\Delta AICc$ level of significance < 2), with model selection
243 facilitated by the utilisation of the dredge function (R Package ‘MuMIn’ [58]). AIC considers
244 the different combinations of explanatory variables (i.e., combinations of season and presence)
245 within a model and as such the scavenging response variables had four possible models: non-
246 interaction season and dingo presence model ($x \sim y + z$), season model ($x \sim y$), dingo presence
247 model ($x \sim z$), and null model ($x \sim 1$).

248 Significance testing (p level of significance < 0.05) was also undertaken using the base model
249 (i.e., non-interaction season and dingo presence model – $x \sim y + z$) to determine which

250 explanatory variables (seasons and/or dingo presence) and/or their levels (summer, autumn,
251 winter, spring) were important in explaining each of the scavenging response variables. To
252 yield additional information from these models, Tukey's honest significance tests were used to
253 determine which seasons were significantly different from one another regarding the modelled
254 scavenging response variable (R Package emmeans) [59].

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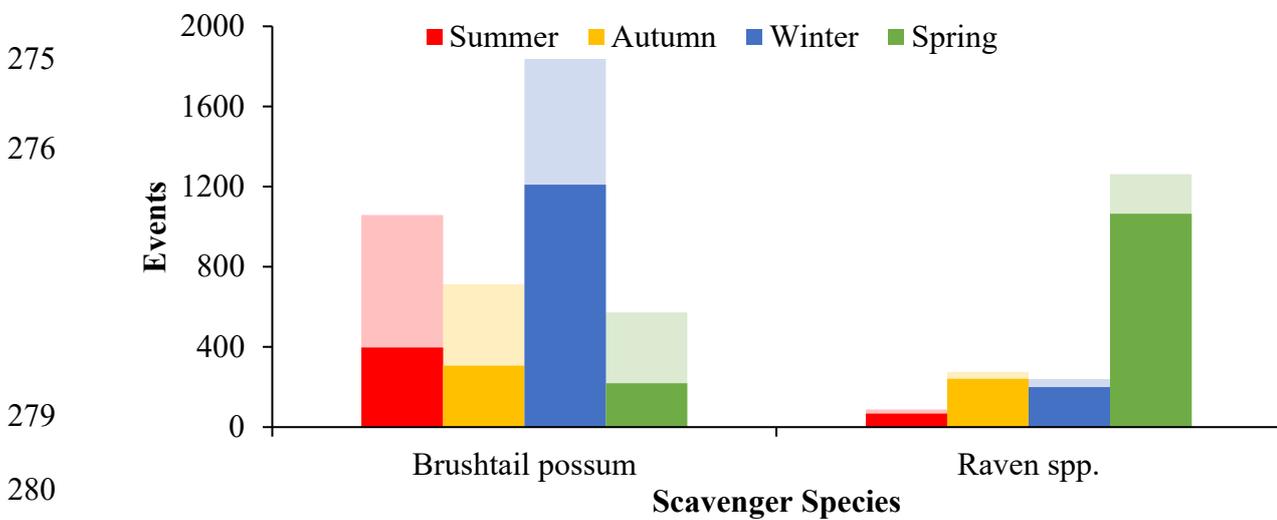
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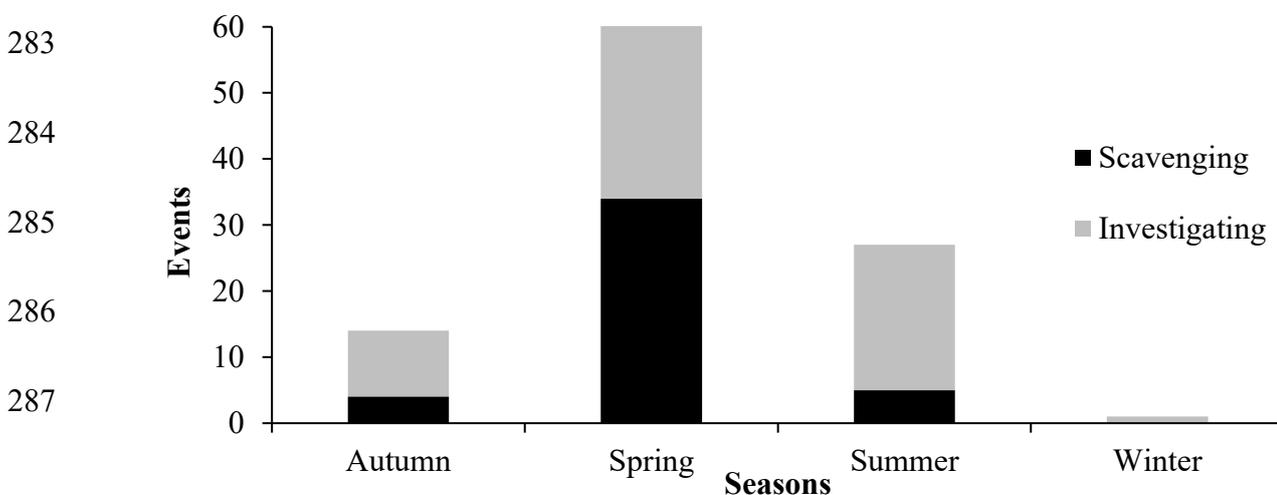
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268 **Results**

269 Of the 60 carcass sites monitored, remote camera data was gathered for 58 – remote camera
 270 data for two sites (one during winter and one during summer) were lost due to theft and camera
 271 failure. In total, there were 247,985 images of raven spp., 225,052 images of brushtail possums,
 272 and 19,772 images of dingoes. These images recorded 4182 distinct events for brushtail
 273 possums, 1866 distinct events for raven spp., and 103 distinct events for dingoes (Figure 2 and
 274 3).

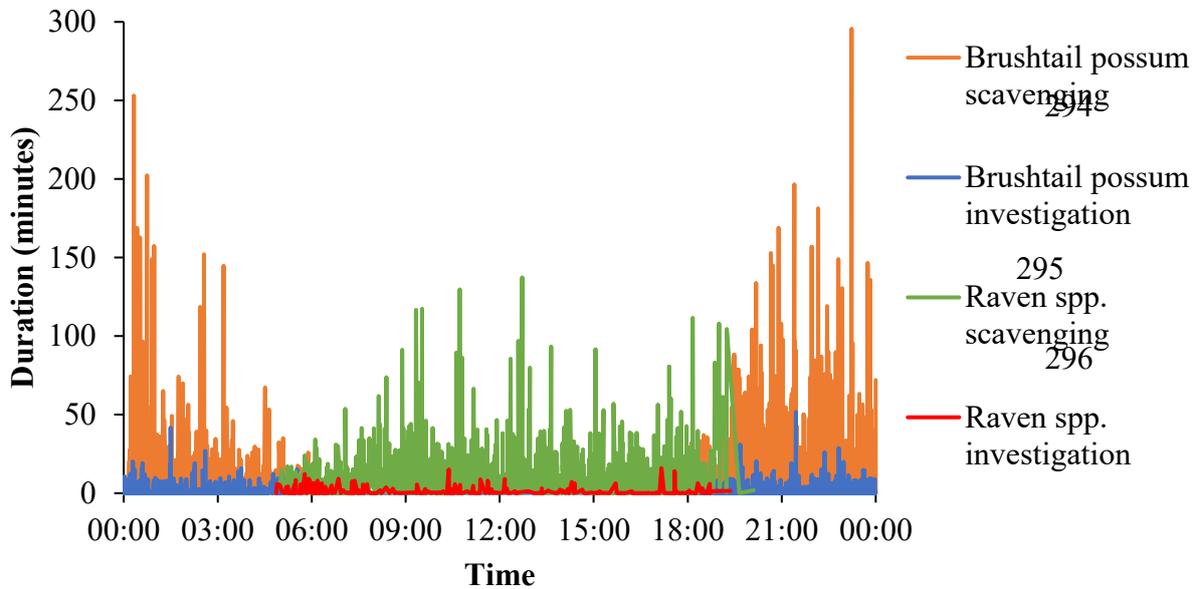


281 **Figure 2.** The total number of investigation events (shaded fill) and scavenging events (solid
 282 fill) each seasons for brushtail possums and raven spp..

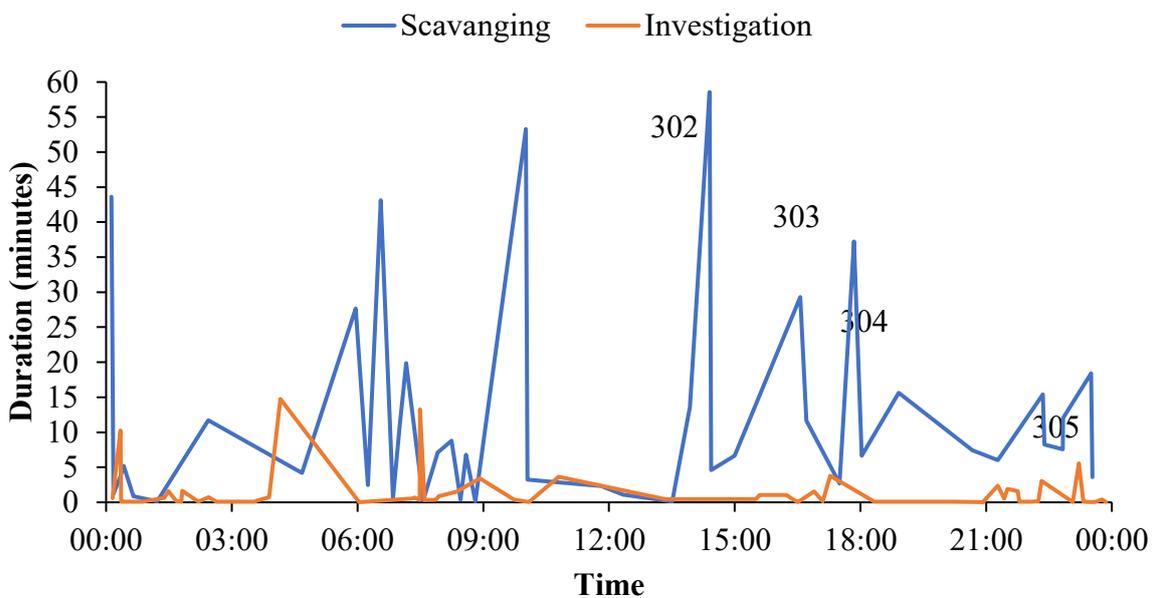


288 **Figure 3.** The total number of dingo investigation events (shaded fill) and scavenging events
 289 (solid fill) each season.

290 The diurnal patterns of scavenging revealed a peak in brushtail possum scavenging frequency
 291 and duration from the early evening until midnight. Whilst raven spp. scavenging frequency
 292 and duration peaked during the middle of the day and late afternoon (Figure 4). Dingoes
 293 seemingly showed no preference for the time of day in which they scavenged (Figure 5).



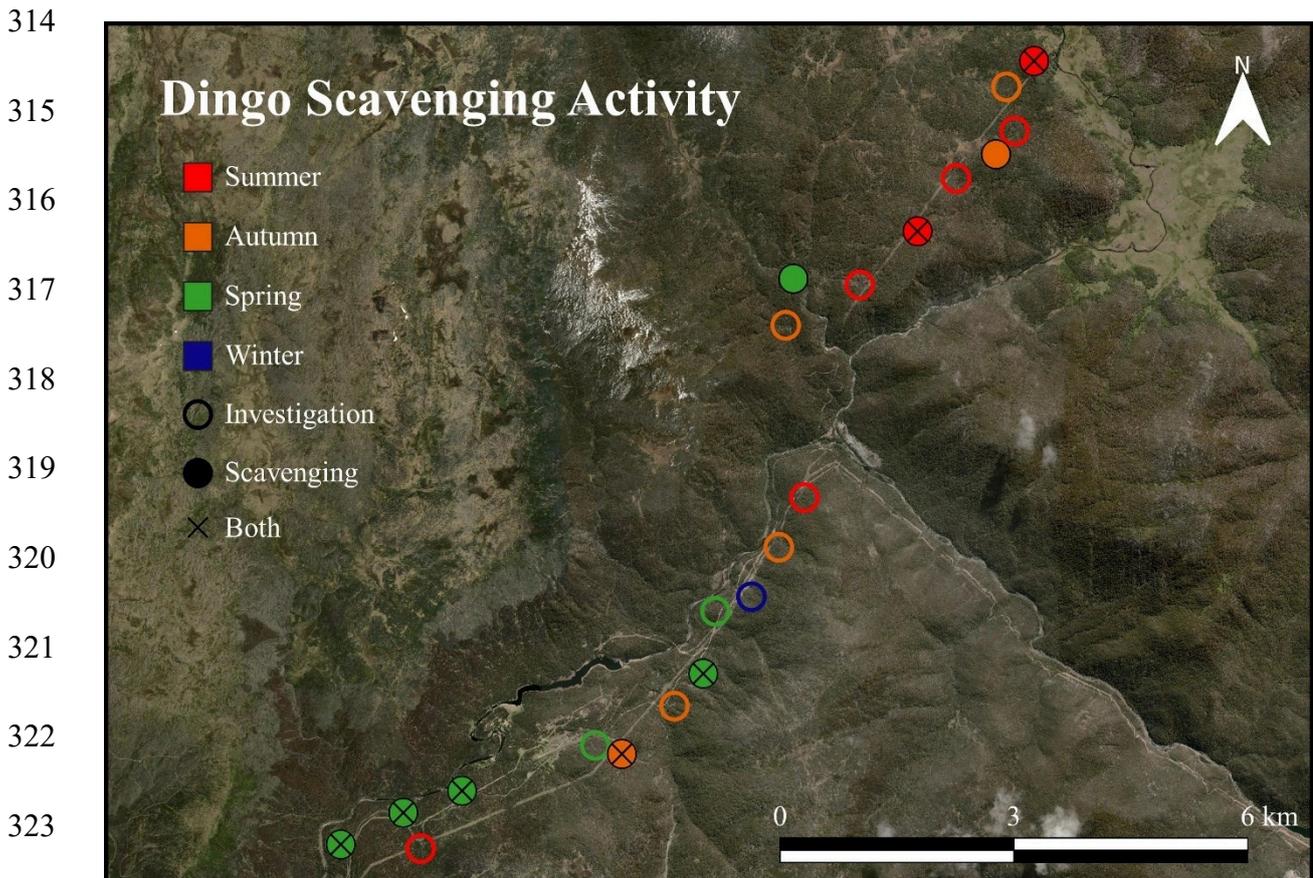
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 299 **Figure 4.** The duration (in minutes) and time (24 hour) of each investigation event and
 300 scavenging event for brushtail possums and raven spp..



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 307 **Figure 5.** The duration (in minutes) and time (24 hour) of each investigation event (orange)
 308 and scavenging event (blue) for dingoes.

309 **Dingo scavenging activity**

310 Two models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in dingo
311 scavenging activity (Table S1). The first model was the seasons model ($\Delta AICc = 0.00$) and the
312 second was the null model ($\Delta AICc = 0.09$). The base model was not significant (Figure 4;
313 Table S2 and S3).



324 **Figure 4.** The carcass sites at which dingoes were recorded to investigate (ring), scavenge
325 (circle), or both investigate and scavenge (X) the monitored carcass during each of the seasons.

326 **Dingo investigation and scavenging duration**

327 One model each was parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in
328 dingo investigation duration and scavenging duration (Table S4). Both these models were the
329 null models ($\Delta AICc = 0.00$; Table S4). The base models were not significant (Table S5 and
330 S6).

331 **Brushtail possum scavenging activity**

332 Two models were parsimoniously competitive ($\Delta\text{AICc} < 2$) in explaining the variation in
333 brushtail possum scavenging activity (Table S7). The first was the season model ($\Delta\text{AICc} =$
334 0.00) and the second the non-interaction season and dingo presence model ($\Delta\text{AICc} = 0.73$).
335 The base model demonstrated that winter had a significant effect on brushtail possum
336 scavenging activity ($p = 0.001$; Table S8). Specifically, brushtail possums were 9.855, 6.117,
337 and 7.029 times more likely to scavenge than investigate a carcass during winter than during
338 autumn ($p = 0.003$), spring ($p = 0.023$), and summer ($p = 0.016$) respectively (Table S9).

339 **Brushtail possum investigation and scavenging duration**

340 Three models and two models were parsimoniously competitive ($\Delta\text{AICc} < 2$) in explaining the
341 variation in brushtail possum investigation event duration and scavenging event duration
342 respectively (Table S10). For both investigation and scavenging duration, the first was the null
343 model ($\Delta\text{AICc} = 0.00$), the second the dingo presence model ($\Delta\text{AICc} = 0.55$ and 0.89), and for
344 investigation event duration the third the non-interaction season and dingo presence model
345 ($\Delta\text{AICc} = 1.56$). The base models were not significant (Table S11 and S12).

346 **Raven spp. scavenging activity**

347 All four models were parsimoniously competitive ($\Delta\text{AICc} < 2$) in explaining the variation in
348 the scavenging activity of raven spp. (Table S13). The first was the null model ($\Delta\text{AICc} = 0.00$),
349 the second the dingo presence model ($\Delta\text{AICc} = 0.10$), the third the season model (ΔAICc
350 $= 1.15$), and the fourth the non-interaction season and dingo presence model ($\Delta\text{AICc} = 1.27$).
351 The base model demonstrated that summer had a significant effect on raven spp. scavenging
352 activity ($p = 0.043$; Table S14).

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355 **Raven spp. investigation and scavenging duration**

356 Two models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in raven
357 spp. investigation event duration (Table S16). The first was the null model ($\Delta AICc = 0.00$) and
358 the second the dingo presence model ($\Delta AICc = 1.92$). The base model was not significant
359 (Table S17 and S18).

360 Two models were parsimoniously competitive ($\Delta AICc < 2$) in explaining the variation in raven
361 spp. scavenging event duration (Table S16). The first was the season model ($\Delta AICc 0.00$) and
362 the second the non-interaction seasons and dingo presence model ($\Delta AICc 2.00$). The base
363 model demonstrated that spring had a significant effect on raven spp. scavenging event duration
364 ($p = < 0.001$; Table S17). Specifically, raven spp. scavenging event duration (in minutes) was
365 1.066 and 1.050 times shorter during spring than during autumn ($p = < 0.001$) and summer (p
366 = 0.029) respectively (Table S18).

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375 **Discussion**

376 Despite being the dominant apex predator and scavenger within the Australian Alps, here, the
377 occurrence of dingoes at carcass sites was found to have no influence on the scavenging
378 activities and behaviours of the brushtail possum or raven spp., contradicting initial hypotheses.
379 Of the 58 carcasses monitored during this study only nine were scavenged on by dingoes
380 (15%). This low rate of dingo scavenging, which peaked in spring and did not occur at all in
381 winter, may allow other scavenger species to access carrion resources more easily (e.g.,
382 brushtail possums and raven spp. – *Chapter 2*), although further manipulative studies would be
383 needed to demonstrate this link. Broadly, these findings also suggest the need to investigate
384 links between dingo population sizes and densities, the occurrence of dingoes at carcass sites,
385 and the severity of their top-down scavenging effects [21].

386 Apex predator density at any spatial or temporal scale within a landscape has been
387 demonstrated to be an important determinant of the extent to which they can exert top-down
388 effects [60-64]. In general, higher predator densities result in greater top-down effects [62, 63,
389 65]. By inference, the lack of top-down scavenging effects exerted by dingoes on brushtail
390 possums and raven spp. in this study could simply be explained by sporadic dingo presence at
391 carcasses (36%), and low dingo scavenging rates (15%), both of which are potentially
392 indicative of low dingo densities. Whilst this is in fact a likely explanation for these results, it
393 does oversimplify density-dependent theories related to the top-down effects of predators.

394 There are many caveats to such theories. For instance, higher predator densities can result in
395 increased inter/intra specific conflict between predators, and consequently, reduce top-down
396 effects on prey species [62, 63]. Such complexities can also occur when considering a
397 predator's density in conjunction with its home range, with both factors sharing an inverse
398 relationship (i.e., as range increases density decreases) [66]. With regards to this study,

399 anecdotal estimates suggest that there were at least three dingo packs recorded at carcass sites
400 during the monitoring, and in fact, an active dingo den was observed within 50 meters of a
401 carcass site at the western end of the transect. However, such observations remain
402 circumstantial without further investigations into dingo population and pack dynamics,
403 densities, home ranges, and the flow-on effects such factors have on other species within a
404 scavenging context and more broadly.

405 Other studies have also described predator densities specifically regarding dingoes and its
406 bearing on the magnitude of their top-down effects [45, 61, 67-69]. One highly relevant study
407 investigated the scavenging dynamics of dingoes, invasive red foxes, and feral cats in an
408 ecosystem similar to that of the Australian Alps [40]. The study, by Forsyth et al. (2014) [40],
409 recorded dingo scavenging rates much higher than those recorded by this study (70% vs 15%).
410 Despite this, Forsyth et al. (2014) found that dingoes contributed very little to carcass biomass
411 loss and attributed this to low dingo densities, a legacy of intensive dingo control in the region.
412 Dingoes are also regularly baited and trapped, in and surrounding, Kosciuszko National Park
413 [70, 71]. This likely knocks down dingo pack size and reduces overall species density in the
414 area [72]. Therefore, similarly to Forsyth et al.'s (2014) proposal that anthropogenically
415 controlled low dingo densities hinder the ability of dingoes to contribute to carcass biomass
416 loss, the same concept may be applied to the dingo's ability to exert top-down effects within
417 scavenger guilds in the Australian Alps. It is apparent that there may be numerous density
418 dependent predator theories that explain the results of this study. Whilst the aims, field design,
419 and results of this study were not geared towards determining which of these theories best
420 explains why dingoes had little effect on the scavenging of brushtail possums and raven spp.,
421 such theories are clearly at the foundations of these findings.

422 In addition to the dingo management regimes undertaken in the Australian Alps, there exists a
423 plethora of invasive species that have also warranted extensive culling programmes to combat

424 their detrimental impacts. The Australian Alps' susceptibility to vertebrate invasion has seen
425 taxa including feral horses (*Equus ferus caballus*), deer (*Cervidae spp.*), feral goats (*Capra*
426 *hircus*), feral pigs (*Sus scrofa*), and rabbits (*Oryctolagus cuniculus*) become established across
427 the landscape since European colonisation. Often, many of the operations undertaken to
428 suppress the populations of such invasive species leave culled animal carcasses *in situ* to decay.
429 As a consequence of these management practises and operations, the Australian Alps may
430 harbour abnormally high rates of carrion biomass.

431 The carcasses used in this study were those of Eastern grey kangaroos (*Macropus giganteus*),
432 a native prey species, and thus, predicted to be highly attractive to dingoes. However, the
433 carcasses of the larger culled ungulates (i.e., horses, deer, and pigs) may be more attractive to
434 scavengers in Australia when compared to the smaller and leaner native prey species i.e.,
435 kangaroos. Therefore, with a likely abundance of available carrion biomass in the Australian
436 Alps, dingoes may be more attracted to and/or selective for the carcasses of larger invasive
437 species than to those of kangaroos. Once again, this study draws parallels to that of Forsyth et
438 al. (2014), as the presence of unknown hunter shot carcasses in the landscape was also
439 presented as an explanation for a lack of dingo scavenging effects [40]. The presence of other
440 unrelated and unknown carcasses within the landscape is an ever-present variable within any
441 experimental field-based scavenging study. However, it is of particular relevance here given
442 the likely high carrion loads present within the Australian Alps.

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447 **Conclusion**

448 It has already been established across much of mainland Australia that dingoes can influence
449 the populations and movements of subordinate species as well as maintain biodiversity and
450 regulate ecosystems [30-37, 45]. It was expected that such apex predator effects would translate
451 into a scavenging context given similar findings for other apex predators globally, specifically
452 canid predators [39]. This, however, was not the case with dingoes seemingly scavenging too
453 little to be able to effectively exert any top-down scavenging effects on either the brushtail
454 possum or raven spp..

455 These results raise several questions regarding the dingo's ability to scavenge and the
456 frequency with which they do so. However, given past observations of relatively frequent dingo
457 scavenging rates [38, 40], these findings are probably indicative of low dingo densities within
458 the landscape. Additionally, given the high number of large bodied invasive species present
459 within the Australian Alps and the culling programmes they warrant, during culling seasons
460 there may be high carrion loads present within the landscape. Carrion saturation within
461 ecosystems may impact the ability of dingoes to exert top-down scavenging effects and/or may
462 have influenced the results of this study if dingoes are more attracted to the carcasses of
463 invasive species than that of native taxa.

464 These conclusions provide considerable scope and future direction for subsequent studies
465 investigating dingo scavenging rates and their consequential scavenging effects and whether
466 they may be dependent on dingo densities or carrion loads. Such concepts and suggested
467 approaches are discussed in the following chapter (*Chapter 4 – Conclusion and future*
468 *directions*).

469

470

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642

Chapter 3 – Supplementary Material

Table 1. The Akaike information criterion (AIC) ranking of each of the models for dingo scavenging activity. Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
<i>1</i>	<i>Season</i>	<i>5</i>	<i>-63.883</i>	<i>136.4</i>	<i>0.00</i>	<i>0.511</i>
<i>2</i>	<i>Null (intercept only)</i>	<i>2</i>	<i>-66.178</i>	<i>136.5</i>	<i>0.09</i>	<i>0.489</i>

Table 2. The base generalised linear model (GLM) for dingo scavenging activity.

Variables	Estimate	Std. error	t value	<i>p</i>
Intercept	-1.109	0.812	-1.366	0.172
Spring	1.292	0.896	1.442	0.149
Summer	-0.713	0.996	-0.716	0.474
Winter	-18.463	1024.000	-0.018	0.986

Table 3. The Tukey’s honest significance test of the base generalised linear model (GLM) for dingo scavenging activity.

Contrast	estimate	SE	df	z.ratio	p.value
Autumn - Spring	-1.292	0.896	Inf	-1.442	0.473
Autumn - Summer	0.713	0.996	Inf	0.716	0.891
Autumn - Winter	18.463	1024.000	Inf	0.018	1.000
Spring - Summer	2.005	0.885	Inf	2.265	0.106
Spring - Winter	19.755	1024.000	Inf	0.019	1.000
Summer - Winter	17.750	1024.001	Inf	0.017	1.000

Table 4. The Akaike information criterion (AIC) ranking of each of the models for dingo investigation event duration (top) and scavenging event duration (bottom). Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
Investigation Event Duration						
<i>1</i>	<i>Null (intercept only)</i>	<i>3</i>	<i>-63.880</i>	<i>134.2</i>	<i>0.00</i>	<i>0.852</i>
2	Season	6	-62.051	137.7	3.5	0.148
Scavenging Event Duration						
<i>1</i>	<i>Null (intercept only)</i>	<i>3</i>	<i>-144.141</i>	<i>294.9</i>	<i>0.00</i>	<i>0.922</i>
2	Season	5	-144.104	299.8	4.93	0.078

Table 5. The base generalised linear model (GLM) for dingo investigation event duration (top) and scavenging event duration (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	p
Investigation Event Duration				
Intercept	0.156	0.450	1.146	0.252
Spring	-0.624	0.527	-1.184	0.236
Summer	0.102	0.543	0.188	0.851
Winter	-1.144	1.493	-0.767	0.443
Scavenging Event Duration				
Intercept	2.340	0.833	2.808	0.005
Spring	-0.209	0.905	-0.231	0.818
Summer	-0.065	1.076	-0.061	0.952

Table 6. The Tukey’s honest significance test of the base generalised linear model (GLM) for dingo investigation event duration (top) and scavenging event duration (bottom). There are no comparisons of dingo scavenging event duration (bottom) during winter because dingoes were not recorded to have scavenged during the winter monitoring period.

Contrast	estimate	SE	df	z.ratio	p.value
Investigation Event Duration					
Autumn - Spring	0.624	0.527	Inf	1.184	0.637
Autumn - Summer	-0.102	0.543	Inf	-0.188	0.998
Autumn - Winter	1.144	1.493	Inf	0.767	0.870
Spring - Summer	-0.726	0.409	Inf	-1.776	0.285
Spring - Winter	0.521	1.450	Inf	0.359	0.984
Summer - Winter	1.247	1.456	Inf	0.857	0.827
Scavenging Event Duration					
Autumn - Spring	0.209	0.905	Inf	0.231	0.971
Autumn - Summer	0.065	1.076	Inf	0.061	0.998
Spring - Summer	-0.144	0.768	Inf	-0.187	0.981

Table 7. The Akaike information criterion (AIC) ranking of each of the models for brushtail possum scavenging activity. Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
<i>1</i>	<i>Season</i>	<i>5</i>	<i>-2537.660</i>	<i>5085.3</i>	<i>0.00</i>	<i>0.574</i>
<i>2</i>	<i>Seasons + Dingo Presence</i>	<i>6</i>	<i>-2537.024</i>	<i>5086.1</i>	<i>0.73</i>	<i>0.398</i>
3	Null (intercept only)	2	-2544.019	5092.0	6.71	0.020
4	Dingo Presence	3	-2543.986	5094.0	8.64	0.008

Table 8. The base generalised linear model (GLM) for brushtail possum scavenging activity.

Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	<i>p</i>
Intercept	-2.224	0.548	-4.060	< 0.001
Spring	0.477	0.633	0.754	0.451
Summer	0.338	0.653	0.518	0.604
Winter	2.288	0.663	3.448	0.001
Dingo Presence	-0.559	0.499	-1.120	0.263

Table 9. The Tukey’s honest significance test of the base generalised linear model (GLM) for brushtail possum scavenging activity. Significant p-values ($p < 0.05$) are denoted in **bold**.

Contrast	estimate	SE	df	z.ratio	p.value
Autumn - Spring	-0.477	0.634	Inf	-0.753	0.876
Autumn - Summer	-0.338	0.654	Inf	-0.517	0.955
Autumn - Winter	-2.288	0.665	Inf	-3.442	0.003
Spring - Summer	0.139	0.638	Inf	0.218	0.996
Spring - Winter	-1.811	0.638	Inf	-2.839	0.023
Summer - Winter	-1.950	0.658	Inf	-2.965	0.016

Table 10. The Akaike information criterion (AIC) ranking of each of the models for brushtail possum investigation event duration (top) and scavenging event duration (bottom). Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
Investigation Event Duration						
1	<i>Null (intercept only)</i>	3	-1947.434	3900.9	0.00	0.404
2	<i>Dingo Presence</i>	4	-1946.703	3901.4	0.55	0.308
3	<i>Season + Dingo Presence</i>	7	-1944.191	3902.4	1.56	0.186
4	Seasons	6	-1945.792	3903.6	2.75	0.102
Scavenging Event Duration						
1	<i>Null (intercept only)</i>	3	-6686.760	13379.5	0.00	0.555
2	<i>Dingo Presence</i>	4	-6686.199	13380.4	0.89	0.356
3	Season	6	-6685.953	13383.9	4.41	0.061
4	Seasons + Dingo Presence	7	-6685.714	13385.5	5.95	0.028

Table 11. The base generalised linear model (GLM) for brushtail possum investigation event duration (top) and scavenging event duration (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	p
Investigation Event Duration				
Intercept	-0.557	0.246	-2.261	0.024
Spring	0.028	0.313	0.090	0.928
Summer	0.421	0.322	1.309	0.191
Winter	0.617	0.322	1.917	0.055
Dingo Presence	0.451	0.249	1.809	0.070
Scavenging Event Duration				
Intercept	1.996	0.383	5.209	< 0.001
Spring	-0.124	0.402	-0.309	0.757
Summer	0.196	0.414	0.474	0.635
Winter	0.156	0.418	0.374	0.709
Dingo Presence	-0.210	0.304	-0.690	0.490

Table 12. The Tukey's honest significance test of the base generalised linear model (GLM) for brushtail possum investigation event duration (top) and scavenging event duration (bottom).

Contrast	estimate	SE	df	z.ratio	p.value
Investigation Event Duration					
Autumn - Spring	-0.028	0.313	Inf	-0.090	1.000
Autumn - Summer	-0.421	0.322	Inf	-1.309	0.557
Autumn - Winter	-0.617	0.322	Inf	-1.917	0.221
Spring - Summer	-0.393	0.324	Inf	-1.213	0.618
Spring - Winter	-0.589	0.328	Inf	-1.796	0.275
Summer - Winter	-0.196	0.336	Inf	-0.584	0.937
Scavenging Event Duration					
Autumn - Spring	0.124	0.402	Inf	0.309	0.990
Autumn - Summer	-0.196	0.414	Inf	-0.474	0.965
Autumn - Winter	-0.156	0.418	Inf	-0.374	0.982
Spring - Summer	-0.320	0.371	Inf	-0.865	0.823
Spring - Winter	-0.280	0.339	Inf	-0.828	0.841
Summer - Winter	0.040	0.355	Inf	0.113	1.000

Table 13. The Akaike information criterion (AIC) ranking of each of the models for raven spp. scavenging activity. Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
<i>1</i>	<i>Null (intercept only)</i>	<i>2</i>	<i>-794.862</i>	<i>1593.7</i>	<i>0.00</i>	<i>0.328</i>
<i>2</i>	<i>Dingo Presence</i>	<i>3</i>	<i>-793.909</i>	<i>1593.8</i>	<i>0.10</i>	<i>0.312</i>
<i>3</i>	<i>Season</i>	<i>5</i>	<i>-792.423</i>	<i>1594.9</i>	<i>1.15</i>	<i>0.185</i>
<i>4</i>	<i>Seasons + Dingo Presence</i>	<i>6</i>	<i>-791.476</i>	<i>1595.0</i>	<i>1.27</i>	<i>0.174</i>

Table 14. The base generalised linear model (GLM) for raven spp. scavenging activity.

Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	<i>p</i>
Intercept	1.820	0.307	5.934	< 0.001
Spring	-0.099	0.337	-0.295	0.768
Summer	-0.865	0.428	-2.021	0.043
Winter	-0.100	0.407	-0.246	0.805
Dingo Presence	0.405	0.287	1.410	0.158

Table 15. The Tukey's honest significance test of the base generalised linear model (GLM) for raven spp. scavenging activity.

Contrast	estimate	SE	df	z.ratio	p.value
Autumn - Spring	0.099	0.337	Inf	0.295	0.991
Autumn - Summer	0.865	0.428	Inf	2.021	0.180
Autumn - Winter	0.100	0.407	Inf	0.246	0.995
Spring - Summer	0.766	0.384	Inf	1.993	0.190
Spring - Winter	0.001	0.362	Inf	0.003	1.000
Summer - Winter	-0.765	0.444	Inf	-1.722	0.312

Table 16. The Akaike information criterion (AIC) ranking of each of the models for raven spp. investigation event duration (top) and scavenging event duration (bottom). Parsimoniously competitive models ($\Delta\text{AICc} < 2.00$) are denoted in *italics*.

No.	Model	df	logLik	AICc	ΔAICc	weight
Investigation Event Duration						
<i>1</i>	<i>Null (intercept only)</i>	<i>3</i>	<i>-267.151</i>	<i>540.4</i>	<i>0.00</i>	<i>0.534</i>
<i>2</i>	<i>Dingo Presence</i>	<i>4</i>	<i>-267.083</i>	<i>542.3</i>	<i>1.92</i>	<i>0.205</i>
3	Season	6	-265.077	542.4	2.06	0.190
4	Seasons + Dingo Presence	7	-265.010	544.4	4.03	0.071
Scavenging Event Duration						
<i>1</i>	<i>Season</i>	<i>6</i>	<i>-5152.495</i>	<i>10317.0</i>	<i>0.00</i>	<i>0.726</i>
<i>2</i>	<i>Season + Dingo Presence</i>	<i>7</i>	<i>-5152.484</i>	<i>10319.0</i>	<i>2.00</i>	<i>0.267</i>
3	Null (intercept only)	3	-5160.473	10327.0	9.92	0.005
4	Dingo Presence	4	-5160.447	10328.9	11.88	0.002

Table 17. The base generalised linear model (GLM) for raven spp. investigation event duration (top) and scavenging event duration (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Variables	Estimate	Std. error	t value	P
Investigation Event Duration				
Intercept	0.456	0.354	1.287	0.198
Spring	-0.246	0.351	-0.702	0.483
Summer	0.156	0.456	0.341	0.733
Winter	-0.733	0.435	-1.683	0.092
Dingo Presence	-0.106	0.286	-0.370	0.711
Scavenging Event Duration				
Intercept	0.065	0.013	4.954	< 0.001
Spring	0.064	0.015	4.312	< 0.001
Summer	0.015	0.018	0.833	0.405
Winter	0.022	0.017	1.247	0.212
Dingo Presence	0.002	0.013	0.144	0.886

Table 18. The Tukey's honest significance test of the base generalised linear model (GLM) for raven spp. investigation event duration (top) and scavenging event duration (bottom). Significant p-values ($p < 0.05$) are denoted in **bold**.

Contrast	estimate	SE	df	z.ratio	p.value
Investigation Event Duration					
Autumn - Spring	0.246	0.351	Inf	0.702	0.897
Autumn - Summer	-0.155	0.456	Inf	-0.341	0.986
Autumn - Winter	0.733	0.435	Inf	1.683	0.333
Spring - Summer	-0.402	0.397	Inf	-1.011	0.743
Spring - Winter	0.487	0.371	Inf	1.312	0.555
Summer - Winter	0.888	0.467	Inf	1.901	0.228
Scavenging Event Duration					
Autumn - Spring	-0.064	0.015	Inf	-4.312	< 0.001
Autumn - Summer	-0.015	0.018	Inf	-0.833	0.839
Autumn - Winter	-0.022	0.017	Inf	-1.247	0.597
Spring - Summer	0.049	0.018	Inf	2.768	0.029
Spring - Winter	0.042	0.017	Inf	2.548	0.053
Summer - Winter	-0.006	0.019	Inf	-0.333	0.987

1 **Chapter 4 – Conclusion and future directions**

2 **Conclusion**

3 The Australian Alps is a unique system where the dynamics of the scavenging guild are
4 dominated by mesoscavengers despite the presence of an apex predator. As expected,
5 scavenging was seasonal but was ultimately dictated by the seasonality in the scavenging
6 activities of the brushtail possum and raven spp. (mesoscavengers). The time to first detection
7 of carcasses was significantly longer during summer and refuted the hypothesis that olfactory
8 cues (signals greatest during warmer months [1]) increase carcass detectability by vertebrate
9 scavengers during summer. A few theories may explain this, the least speculative of which
10 regards the brushtail possums and raven spp. abilities to rapidly detect carcasses. During
11 autumn, winter, and spring, 93% of the first detections of a carcass were by either brushtail
12 possums or raven spp., but during summer they only accounted for 57%. This reduction in
13 search effort by the brushtail possum and raven spp. during summer, when there is likely an
14 abundance of other available food sources [2-4], may have alone increased the time to first
15 detection of carcasses for the collective scavenger guild. Increased invertebrate scavenging
16 rates during summer have been demonstrated to accelerate carcass decomposition, and in some
17 cases even outcompete vertebrate scavengers' contribution to carrion biomass loss [5-7]. This
18 may also explain why the time to first detection of carcasses was so long during summer as
19 detection by vertebrates becomes increasingly difficult during the later stages of
20 decomposition, and indeed less attractive [8, 9].

21 Collectively, the scavenger guild was significantly more likely to scavenge than investigate
22 carcasses during both spring and winter. For spring, this was driven by the raven spp. breeding
23 season, during which it was inferred that raven spp. became highly dependent on carcasses to
24 meet increased energetic requirements, as well as for nest construction (i.e., hair/fur), and

25 feeding chicks [10, 11]. The highly frequent but brief raven spp. scavenging events associated
26 with the raven spp. breeding season i.e., frequent trips back and forth between the nest and
27 resources [10], also potentially resulted in significantly shorter scavenging events during spring
28 for the collective scavenger guild. With regards to winter, increased rates of scavenging vs
29 investigation of carcasses were consistent with results observed globally, especially so in
30 northern hemisphere systems, where rates of scavenging increase during winter in response to
31 a lack of other available food sources [2-4]. However, here, this winter-time trend was
32 completely dominated by an increase in the scavenging rates of brushtail possums which
33 scavenged three times more often than during summer and accounted for 81% of all recorded
34 events during winter. The rates of scavenging observed by possums during winter, and indeed
35 during all seasons, provide evidence for considerable carnivory in the brushtail possum.

36 Despite the presence of the dingo, a dominant scavenging predator [12-15], within the Australia
37 alpine scavenger guild, the scavenging rates, and behaviours of the brushtail possum and raven
38 spp. were unaffected. Together the results of this study demonstrate a seasonal dynamic in the
39 Australian Alps where scavenging is dominated by low ranking mesoscavengers in the notable
40 absence of the dominant scavenging predator. With this in mind, a conceptual framework can
41 be developed, which is of continental significance in Australia given that brushtail possums are
42 the most widespread Australian marsupial and that both raven spp. are also relatively abundant
43 throughout southern-east Australia. Here, autumn represents a relative 'norm', where
44 scavenging is not completely dominated by the brushtail possum and raven spp.. Notably, the
45 highest rates of scavenging were also recorded for many of the other scavenger's species during
46 autumn. Summer recorded the least amount of scavenging for the collective scavenger guild,
47 especially so regarding brushtail possums and raven spp., and this was likely due to an
48 abundance of other available food sources. Winter was completely dominated by a profound
49 increase in the scavenging activities of brushtail possums as they supplemented their diet in the

50 absence of other available food sources. Finally, the scavenging dynamics of spring were best
51 explained by its overlap with the raven spp. breeding season which resulted in just over 1000
52 frequent but short raven spp. scavenging events. Additionally, dingo scavenging peaked during
53 spring but had no effect on the scavenging of either raven spp. or brushtail possums (Figure 1).



54 **Figure 1.** Conceptual diagram indicating the proportion (i.e., size of wedge) of total scavenging
55 events attributed to each season (yellow/top = summer, blue/right = winter, orange/bottom =
56 autumn, and green/left = spring). Within each season wedge the scavenging dynamics of the
57 three dominant mesoscarvengers (brush-tail possum and raven spp.), as well as other minor
58 scavengers, are shown, where the relative size of the arrow is indicative of interaction strength.

59 **Future directions**

60 Foremost, this study exemplifies the importance of the seasons to ecological processes linked
61 to scavenging. The direction in which seasonal effects impacted scavenging was not always as
62 predicted. These unexpected influences highlight the need for seasonally replicated
63 experimental approaches in field-based scavenging research. This was a primary motivation
64 for undertaking this study as it was obvious that much of the literature describing scavenging
65 ecology in the field inadequately accounted for seasonal effects, underestimating their potential
66 to shape ecosystem scavenging dynamics.

67 In general, the methods used herein can be utilised to monitor scavenging dynamics
68 surrounding carrion within any given ecosystem globally, and even be further developed to
69 improve field-based designs and yield additional ecological information. Such improvements
70 could include utilising an increased altitudinal gradient that may detect elevational differences
71 in scavenging dynamics missed by this study given the small altitudinal gradient used (500m).
72 This is especially important in scavenging studies elsewhere globally that experience much
73 higher altitudes than those present in Australia. Increased sampling, including through the use
74 of multiple transects, along with replicated studies over multiple years may also help to account
75 for within site and annual variability [16].

76 A limiting factor of this study was the absence of measures for carrion biomass loss, and thus,
77 species-specific contributions to biomass loss through scavenging were not estimated. Such
78 findings would have been particularly interesting here given that the contributions of the
79 brushtail possums and raven spp. to biomass loss may have been minimal when compared to
80 larger scavengers (e.g., dingo, feral pig, wedge-tail eagle) capable of consuming greater
81 quantities of carrion, despite scavenging less frequently. This means there is considerable scope
82 to further our understanding of species-specific scavenging dynamics within the Australian

83 Alps by assessing the contributions of different scavengers to carrion persistence within the
84 landscape.

85 Overall, the widescale carnivory of brushtail possums observed in this study is highly novel
86 given that brushtail possums are commonly considered generalist herbivores/folivores [17-21].

87 Whilst brushtail possums have been known to on occasion predate insects, depredate birds'
88 nests, and scavenge on carrion [22, 23], generally their diet is known to consist of leaves,
89 flowers, and fruit (commonly *Eucalyptus* and *Acacia* species) [17, 21, 24]. In light of the
90 findings here and their commonly cited dietary classifications, it is recommended that further
91 dietary studies of the brushtail possum are undertaken which may potentially results in the
92 reclassification of their dietary status. An experimental study that utilises scat analyses, a
93 commonly used method in the literature [17, 24], could be particularly insightful here. Such an
94 approach could compare the diets of brushtail possums (determined by scat analyses) that have
95 access to carrion (experimentally placed carcasses) vs those that do not, and this may determine
96 the extent to which carrion can supplement or even substitute other common food sources.

97 Seemingly, the raven spp. within the Australian Alps were somewhat dependent on carrion as
98 both a source of energy and nesting material during the breeding season. Very little work has
99 been done to determine how carrion biomass within an ecosystem (year to year) may affect the
100 breeding success of facultative scavengers [7]. This is an area that demands future attention
101 given the results of this study and also those elsewhere regarding raven spp. [10]. A similar
102 approach to that discussed above regarding brushtail possums (i.e., access vs exclusion to
103 carrion) could also be utilised here for comparisons of raven spp. breeding success across
104 multiple breeding seasons.

105 Of global relevance is the seemingly low rates of dingo scavenging in the Australian Alps and
106 the consequential lack of top-down scavenging effects exerted by an apex predator surrounding
107 carrion. This raises several questions regarding how dominant scavenging predators affect

108 scavenging dynamics under different scenarios. For this reason, specifically, there is a need for
109 simultaneous comparisons of scavenging effects exerted by dominant scavenging predators in
110 ecosystems where the predators are not subject to control vs ecosystems where predators are
111 anthropogenically controlled or are already extirpated. Such comparisons would be highly
112 valuable globally given the unique and varying circumstances under which apex predators
113 exists. These findings would also assist in determining whether the lack of scavenging and
114 scavenging effects by the dingo in the Australian Alps, is unique, or part of a more global trend,
115 where the abilities of apex predators to control and regulate food web dynamics within
116 ecosystems are being negatively impacted by anthropogenic activities [25-30].

117 The density dependent predator theories that were likely important to the lack of top-down
118 scavenging effects exerted by dingoes in this study, provide a strong foundation for such
119 comparative predator studies recommended above [31, 32]. Indeed, such studies have already
120 been undertaken in Tasmania where a reduction in the density of Tasmanian devils resulted in
121 a mesoscavenger release of forest ravens and feral cats [33, 34]. These studies exemplify such
122 an approach that can link density dependent predator theories with scavenging and serve as
123 relevant benchmark studies. However, further refinements (i.e., additional seasonal
124 replications) are required and this could be specifically undertaken within the Australian Alps
125 given the routine dingo control that take place within and surrounding Kosciuszko National
126 Park [35, 36]. Such programmes have been demonstrated to negatively impact dingo density
127 and pack size [37]. Therefore, by monitoring dingo density (i.e., camera trap grid) as a function
128 of anthropogenic control, in tandem with scavenging dynamics surrounding carrion, the flow-
129 on effects of dingo densities on dingo scavenging rates and top-down scavenging effects may
130 be uncovered. In addition, more informative measures of dingo top-down scavenging effects
131 other than that used in this study (i.e., presence or absence of dingoes at carcass sites) should
132 be utilised in future studies to characterise more realistic effects on mesoscavengers. This could

133 include the number of visitation to a carcass by dingoes, the time spent scavenging at a carcass,
134 and/or time since last dingo visitation to a carcass.

135 These types of approaches could be further supplemented by a robust quantification of the
136 relative contribution of scavenging vs predation in the diet of the dingo. The most recent
137 continental review of the dingo diet suggested that 66% consisted of mammals, 22% birds, and
138 11% reptiles [38]. However, the dataset used did not provide the capacity to quantify what
139 percentage of that was scavenged vs predated, and indeed, Doherty et al. (2019) highlighted
140 this as an area of further study. A greater understanding of the dingo's diet would be a good
141 first step towards determining how important an energy pathway scavenging is to dingoes, and
142 thus, how important a species they are within their scavenging guilds and the likely extent to
143 which they can exert top-down scavenging effects [39].

144 Lastly, the Australian Alps likely harbours high loads of carrion biomass due to the widespread
145 presence of larger bodied invasive species. The extensive management programs they warrant
146 can also pulse carrion biomass loads within the ecosystem during culling seasons. This raises
147 two points: *firstly*, that estimates of carrion biomass within the Australian Alps, and other
148 ecosystems more broadly, is a field that warrants further attention to determine how over
149 availability or scarcity of carrion affects scavenging dynamics. *Secondly*, it highlights a
150 relatively novel concept that has seldom been explored and experimentally quantified, the
151 proportion to which scavengers utilise the carcasses of native species vs invasive species [40,
152 41]. Research in both these areas could yield particularly insightful findings that could explain
153 how scavengers react to a variety of unpredictable, abundant, and/or scarce carrion sources,
154 and thus, further develop concepts in the scavenging ecology field.

155

156

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